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| 1 | Policy and legislative barriers to close water-related loops in innovative small water and |
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| 2 | wastewater systems in Europe: a critical analysis |
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| 22 | ABSTRACT |
| 23 | Water supply and reuse through non-conventional water resources can significantly decrease |
| 24 | the stress on natural water resources. Decentralized systems can help not only to alleviate |
| 25 | issues of water security in arid areas, but also to create a sustainable framework within a |

26 circular economy. Although these small-scale innovative technologies are able to achieve 27 ready-to-use, high quality of recovered/treated water on-site, the loop cannot be closed in most cases due to legislative barriers. Similarly, the end-use of sewage sludge after treatment 28 in decentralized systems still lacks specific regulations that limit its valorization. This work 29 analyzes the current policy and legislation related to water supply, wastewater treatment, 30 water reuse, and resource valorization within the context of decentralized state-of-the-art 31 technologies applied in rural areas. The drawbacks in the current EU legislation that set 32 33 barriers to close water-related loops in European countries are highlighted. A regulatory fitness check is applied to each type of loop to identify the key factors to accomplish the 34 35 legislative compliance, and financing pathways are further evaluated at the EU level. As a 36 possible solution, further development of an innovation deal approach is recommended to address the environmental, regulatory, and financial gaps in water management through an 37 integrated framework, providing ad-hoc policies and prescriptions for the sustainable reuse of 38 39 all water resources.

Keywords: environmental policy; innovation; non-conventional water resource; rural area;
sustainability; wastewater reuse

42 Word count: 7714 (including the text, tables and figures)

43

44 1. Introduction

The depletion of natural resources at a fast rate leads to a transition of the current society to re-evaluate these resources in a sustainable manner. Water is a fundamental resource to

- 47 sustain life and irregularly distributed both spatially and temporally; furthermore,
- 48 anthropogenic activities continuously contaminate the limited water reserves (Voulvoulis,

49 2018). Consequently, water sustainability is among the most discussed sustainability issues in

50 the last years where every applicable sustainability principle has been adopted to water from

reuse to recycle (Sodiq et al., 2019). These issues are increasingly discussed by the European 51 52 Commission as Europe has become more and more vulnerable to water shortages and to the social, economic, and environmental impacts deriving from increasing demand and global 53 climate change in recent times. In fact, in several areas of Europe, there's a critical issue 54 regarding the balance between the demand and availability of water. As a result, the reduction 55 of available water resources has been followed by a deterioration in the quality of water 56 caused by poor dilution of pollutants. Consequently, EC sets increasingly ambitious 57 objectives to cope with these environmental pressures (European Targets2020, 2020; 58 European Targets2030, 2020). 59

The circular economy concept has been developed to overcome the problems of a linear 60 61 economy 'take-make-use-dispose' model and found great application areas in the water sector to preserve the availability of water (Sodiq et al., 2019; Voulvoulis, 2018). The linear 62 economy aims to treat waste streams involving a potential risk to the receiving environment, 63 while recovery/reuse strategies belong to the circular economy concept (Robles et al., 2020). 64 This transition triggered innovative technologies/processes for efficient water utilization, 65 finding alternative water sources, and closing the water-related loops to balance water demand 66 and supply (Peng et al., 2019). 67

68 The design and operation of a water supply and treatment system should ensure the 69 sustainability of the technology considering the water-energy-food-ecosystem (WEFE) nexus in urban and rural planning (Vakilifard et al., 2018). Land boundaries are of great 70 71 concern during the implementation of these technologies in terms of the governance of environmental resources. There are two major concepts used by policymakers, researchers, 72 73 national administrations, and international organizations for the characterization of settlements as rural and urban areas. The most commonly used method, for the identification 74 of areas, was developed by the Organisation for Economic Cooperation and Development 75

(OECD) and is based on population density (Brezzi et al., 2011). The OECD method
classifies areas with a population density below 150 inhabitants/km² as rural. Moreover, the
method also addressed the predominantly urban, intermediate, and predominantly rural
regions when the share of the population is below 15%, between 15%-50%, and higher than
50%, respectively.

According to the World Health Organisation (WHO, 2016), in the pan-European region, 81 approximately 264 million people lived in rural areas in 2015 and 40% of the population had 82 no access to wastewater collection and treatment systems. For instance, 15-20% of the 83 population in Estonia is not supplied with centralized sewer systems due to dispersed rural 84 85 settlement (Spin Project, http://www.spin-project.eu). At this point, small-scale decentralized 86 collection and treatment systems can bring not only a long-term solution for small and rural communities, but is also reliable, flexible, and cost-effective. Furthermore, adopting 87 decentralized solutions may advance conditions of sustainability and resilience in water 88 management (Leigh and Lee, 2019). 89

Innovative technologies and concepts for water and wastewater systems already exist, but 90 they have been mostly implemented in pilot/demonstrative projects so far, mainly as a result 91 92 of the institutional barriers they face (Trapp et al., 2017). Although the technological, 93 ecological, and economical sustainability of decentralized water/wastewater treatment systems are often promising; the adoption of decentralized systems often fails to go beyond. 94 In fact, crucial dynamics relating to how the water sector can shift towards decentralized 95 96 infrastructure are not well-understood where technological approaches are not sufficient to promote more sustainable systems. There is no doubt that an innovation in the water 97 infrastructure takes place both at the technical and at the institutional or organisational level. 98 In order to replicate these technologies and close the water-related loops, these solutions 99 should be evaluated under the legislative/policy frameworks at the macro-scale. In most cases, 100

101 the European Union (EU) legislations are the starting-points for the EU Member States and accession countries. At this point, we analyzed policy and legislative barriers that hinder the 102 adoption of alternative decentralized systems by seeking answers to the following questions: 103 What is the state of the art on small-scale water cycle facilities in Europe? Which legislations 104 105 are related to water loops at the EU level? What kind of barriers should be dealt with to close water-related loops? What are the key considerations while developing projects which aim to 106 close the water loops? Is there any difference between rural and urban environments regarding 107 108 the limitations? What are the possible solutions for overcoming the existing barriers? To address these questions, this work critically assesses the fitness within relevant EU directives, 109 110 on-going policy initiatives, and minimum requested quality standards, regulatory, and 111 financing frameworks.

112 2. Technology readiness levels of solutions to close water loops

Technologies on non-conventional water resources are becoming fundamental contributors in 113 the water loop such as desalination of seawater and brackish water, rainwater harvesting, 114 atmospheric water harvesting, and wastewater reuse. Today, we have technologies (e.g. 115 desalination of seawater and highly brackish groundwater via osmosis or distillation; 116 117 rainwater harvesting systems by means of micro/macro catchment areas or fog harvesting, 118 etc.), valorizing these water sources at high technology readiness level (TRL) which can be implemented to partially alleviate water scarcity in rural areas where renewable water 119 resources are extremely scarce (Imteaz et al., 2015; Qadir et al., 2007). In addition to potable 120 121 water supply through non-conventional water resources, treated wastewater can be used for different purposes such as irrigation in the agricultural fields or parks, restoration of water 122 bodies and wetlands, recharging in the aquifer for storage, etc. Over the last decade, many 123 researchers have studied various decentralized options for wastewater treatment and reuse 124 (Lijó et al., 2017), such as membrane bioreactor (MBR) (Tai et al., 2014), constructed 125

wetland (CW) (Nivala et al., 2019; Wu et al., 2015), or integrated systems (e.g., bioreactor + 126 CW) (Tanner et al., 2012). In smalls-scale decentralized systems (onsite systems, population 127 1-40), the resources recovered consist of water and nutrients, but at least one option exists for 128 energy recovery. Meanwhile, medium-scale (satellite) facilities can serve 20-47,000 129 inhabitants with a minimum capacity of 8 m³/d and maximum flow of 20,000 m³/d (Diaz-130 Elsayed et al., 2019). Decentralized wastewater treatment systems favor water recycling and 131 reuse in the proximity of their location, while other resources can be readily recycled as bio-132 133 energy and nutrients (Capodaglio, 2017). Considering sludge processing, co-treatment with biowaste may offer a promising solution either by decentralized anaerobic digesters (Thiriet et 134 135 al., 2020) or composting systems (Panaretou et al., 2019). Some examples from Europe are 136 summarized in the e-Supplementary file to highlight the most commonly used decentralized technologies at high TRLs. For instance, Meuler et al. (2008) used both decentralized 137 membrane bioreactors for the reuse of greywater and rainwater harvesting systems to produce 138 water for irrigation or as service water in households. The authors further confirmed their 139 findings' compliance with the national (German) requirements for treated effluent reuse. In 140 another study, Yan et al. (2018) set up a rainwater harvesting system in an office building on 141 142 the University of Exeter's Streatham campus with around 300 occupants. Although the 143 rainwater harvesting system aimed to reduce water consumption in the toilet flushing, the system enabled to get water with a quality met with the criteria for the potable water in the 144 UK which can also be used for washing the building and drinking. Since there is no regulation 145 146 in the UK for such systems, there is a code for practice (BS 8515:2009) specifically for rainwater harvesting covers design, installation, and maintenance of the system, water quality, 147 148 and risk management and rainwater collection systems encouraged in The Code for Sustainable Homes. In Noorderhoek (Netherlands), a biogas plant serving 232 apartments is 149 under operation since 2007 (Bautista Angeli et al., 2018). An UASB reactor (2.5-7 m³) is 150

operated with blackwater, kitchen waste, and greywater under mesophilic conditions and 151 produces 13.8-12.2 m³ CH₄/cap/year which is equal to 133-148 kWh/cap/year heat with CHP 152 unit. The national regulations such as "Regulation designating sustainable energy production 153 categories (enforced 01.10.2014)" for definition of technologies and regulation for subsidies 154 and "Renewable Energy Production Incentive Scheme (enforced 16.10.2007, last recast 155 01.04.2017) for the market price of the electricity and biomethane must be followed for the 156 renewable energy projects in the Netherlands (Hermann et al., 2019). In the study of Starkl et 157 al. (2007), a policy-oriented approach was followed to develop an integrated assessment for 158 rural wastewater management in Austria based on a separation of the wastewater into its 159 constituent parts using various technologies. It was concluded that the co-treatment of black-160 161 water in a regional biogas plant would be technically feasible, but it is not supported by regulations in Austria as well as in my EU member countries. In fact, limitations in the ability 162 of governance structures to adapt was stated as one reason for the stagnation in the 163 implementation of novel water systems in Germany (Schramm et al., 2018). Although these 164 technologies bring innovative solutions to decrease water and energy stress in the regions they 165 are applied, most of them may face legislative obstacles for further reuse and/or valorization 166 167 of resources as analyzed in the following sections.

168 **3. Legislative framework and barriers**

Following the afore-mentioned water stress and possible decentralized solutions, the enabling environment was initially analyzed by checking whether the relevant policies support or hinder the implementation of small-scale decentralized collection and treatment systems when inputs (e.g. water categories) are considered to produce and reuse different outputs (e.g. reclaimed water and recovered materials and potentially marketable products). **Table 1** shows the relevant legislation, policies, and guidance for the input and output of water-related loops.

- 175 Fitness check was specifically assessed within the following main directives. A summary of
- these directives are given in the **e-Supplementary file**.
- **Table 1.** Fitness check-in international policy/regulatory/guidance framework.

| Directive / Regulation / Decision / Recommendation / Guidelines | Relevant input | Relevant output |
|---------------------------------------------------------------------|------------------------|----------------------|
| | Relevant input | Kelevant output |
| European Parliament, 2020, Regulation (EU) 2020/741 of the | | |
| European Parliament and of the Council of 25 May 2020 on minimum | | |
| requirements for water reuse | - | |
| European Commission Council Directive 91/271/EEC (amendment | | |
| 98/15/EC) (Urban Waste Water Treatment Directive) and ongoing | r b | |
| revision COM (2017) 749) | | |
| World Health Organization Guidelines for the Safe Use of | - - | |
| Wastewater, Excreta and Greywater (2006) | Maria 1/Decordia | WZ a tana Cara |
| ISO/TC 282 (2015)/ ISO16075-2:2015 (2015) | Municipal/Domestic | Water for |
| EN 12566- Small wastewater treatment systems for up to 50 PT (Parts | wastewater | irrigation reuse |
| 1-7) | | |
| European Commission, Sewage Sludge Directive 86/278/EEC, | | |
| followed by EC 219/2009a) | , | |
| STRUBIAS Technical Proposal (JRC Science for Policy report, | - | |
| 2019) | , | |
| | | |
| EC report on Digestate and compost as Fertilizers (Corden et al., | , | |
| 2019) | | |
| European Commission, 178/2002 on procedures in matters of food | Water for irrigation, | Crops (for Food |
| safety (2002) | compost seawater and | and industrial |
| European Commission, 2006a (EC 1881/2006 on maximum levels | domestic wastewater | uses), Salts from |
| for certain contaminants in foodstuffs) | domestic wastewater | brine |
| EC Best Environmental Management Practice in The Tourism Sector | • | Rainwater for: |
| (Styles et al., 2013) | Di | Irrigation, drinking |
| Environment Agency, 2010 (Harvesting Rainwater for Domestic | Rainwater | water, and |
| Uses: An Information Guide) | | domestic uses |
| European Commission, 1998 (European Commission, 83/1998 EC | 1 | |
| Drinking Water Directive) and amendment (European Commission, | | |
| 2015a) | | |
| European Commission, 2017 (Proposal EC COM 753/2017 on the | - | |
| quality of water intended for human consumption) | | |
| | Rainwater, water vapor | Drinking water |
| World Health Organization, Guidelines for Drinking-Water Quality | 1 | - |
| (GDWG) (2017) | - | |
| World Health Organization, 2011 (Small-scale drinking water | | |
| supplies in the pan-European region) and World Health Organization, | , | |
| 2012 (Water safety planning for small community water supplies) | | |
| European Commission, 2006b (EC 118/2006 - The Groundwater | • | |
| Directive-GD) | | |
| European Commission, 2000 (2000/60/CE - Water Framework | Rainwater, stormwater | Water for aquifer |
| Directive-WFD) | runoff | recharge/storage |
| European Commission, 2014 (EU 80/2014 on | | 0 0 |
| the protection of groundwater against pollution and deterioration) | | |
| European Commission, 2003 (EC 2003/2003 Fertilizer Regulation) | | |
| and following European Commission, 2003 (EC 2003) 2003 and European | Municipal/Domestic | Recovered |
| Commission, 2019 | wastewater | fertilizers |
| | | |
| European Commission, 2018 (EC 2001/2018 on the promotion of the | | |
| use of energy from renewable sources) | | 1 |
| | | |
| CEN - EN 16726, 2015 (European standard on the quality of gas of | - | Biogas for biofuel |
| the H category) | wastewater | Biogas for biofuel |
| | wastewater | Biogas for biofuel |

3.1. Water reclamation and sanitation for water safety

When considering the implementation of any solution that aims at ensuring sustainable 179 management of water and sanitation (Sustainable Development Goal-SDG 6) in the EU, the 180 fitness check with the Water Framework Directive (WFD) 2000/60/EC cannot be overlooked. 181 In the WFD, the use of reclaimed water is considered as a means of increasing water 182 availability while ensuring a good quality status of water resources. Specifically, the Directive 183 (Annex VI(x)) refers to 'efficiency, reuse measures, and water-saving techniques for 184 irrigation' to help to achieve good environmental status. In this perspective, the 185 implementation of small-scale decentralized technologies could contribute to tackling the 186 187 problem of reaching a good status in Europe, as already highlighted in the <u>Regulatory Fitness</u> 188 and Performance program evaluation (REFIT) of the WFD by the European Environment Agency (European Commission, 2019b). In this perspective, the WFD (Art.11(3-f)) allows to 189 190 artificially recharge the groundwater bodies with water that "...may be derived from any surface water or groundwater...", after the necessary authorization. Thus, no clear constraint 191 on the use of specific water sources is stated, as long as the water used does not compromise 192 the achievement of the environmental objectives for a good water status. 193 194 Furthermore, no explicit permission or prevention is detected for drinking water production 195 from rainwater, as the WFD referred only to conventional water bodies as sources for 196 drinking water production (Art.7(1) and (2)). No specific barriers are identified even when considering the compliance with the Directive 197 198 2006/118/EC, "Groundwater Directive" (GD). However, a focus on national/local legislation should be further assessed to evaluate how monitoring strategies are carried out in different 199 200 countries to ensure the safety of aquifer recharge regarding contamination (e.g. pesticides) by

201 the stormwater runoff.

Compliance with the Council Directive 91/271/EEC "Urban Waste Water Treatment 202 203 Directive" (UWWTD) is crucial for the implementation of small-scale decentralized systems, when reclaimed water production from wastewater is involved. In this regard, the UWWTD 204 (Art.12(1)), promotes the reuse of "treated wastewater...whenever appropriate", as long as it 205 is not prohibited by other EU legislative instruments and does not implicate environmental 206 deterioration. Therefore, no limitations are detected for treated wastewater reuse when quality 207 standards are achieved. To comply with the UWWD's requirements, the priority was given by 208 209 EU member states to urban areas where huge investments in wastewater collection and treatment systems took place. This situation may lead to rural areas to take a backseat. 210 211 Furthermore, the WWTD (Art. 14(1)) promotes the reuse of sludge from WWT "...whenever 212 appropriate...". Although this generic statement does not define specific prescriptions for reuse, it does not forbid the implementation of technologies whose objective is the treatment 213 and the subsequent reuse of sewage sludge. 214 The recent (December 2019) REFIT of the UWWTD (European Commission, 2019c) 215 highlighted that further efforts are still necessary to reach the full compliance with the 216 WWTD in terms of the collection, secondary and stricter treatment applied to wastewater 217 (compliance decreased from 98.4% to 94.7%, from 91.9% to 88.7% and from 87.9% to 84.5% 218 219 respectively) (European Commission, COM (2017) 749 final). During the consultation period of the EU, the consortium of the Innovation Action project of 220 HYDROUSA focused its evaluation of the UWWTD at the challenge of individual or 221 222 appropriate systems. The consortium commented that "Despite the generally high level of implementation of the UWWTD, a number of challenges remain, including the need for 223 further investments in the wastewater sector to increase or maintain implementation, 224 operating costs optimisation, individual or appropriate systems (IAS), stormwater overflows, 225 as well as improving coherence with other European Union water policy." And concluded 226

| 227 | that "IAS should be framed by specific regulations in the future. Specifically, the |
|-----|-----------------------------------------------------------------------------------------------|
| 228 | requirements for designing, constructing, and maintaining IAS must be defined and |
| 229 | environmental protection must be ensured on the same level as a collecting system followed |
| 230 | by centralised wastewater treatment." |
| 231 | It has to be remarked that when considering IAS, small individual wastewater treatment plants |
| 232 | and septic/storage/holding tanks should be provided (European Commission, 2015b) for small |
| 233 | agglomerate (up to 50 PT) or isolated houses. Specific prescriptions on sizing criteria and |
| 234 | operation of septic tanks, prefabricated treatment units, and tertiary treatment should be |
| 235 | followed according to standard EN12566 1-7:2016 "Small wastewater treatment systems for |
| 236 | up to 50 PT". |
| 237 | To increase the compliance, small-scale decentralized systems might contribute to support the |
| 238 | initiatives of the European Commission (European Commission, COM (2017) 749 final) as |
| 239 | follows: |
| 240 | • improve the sludge quality and recovery; |
| 241 | • minimize the consequences of the stormwater overflows pollution; |
| 242 | • increase the treated wastewater reuse, while ensuring appropriate water quality; |
| 243 | • reduce the energy demand of sanitation systems, using (when possible) energy from |
| 244 | renewable resources at the treatment plant (e.g. biogas). |
| 245 | In the context of reclaimed water reuse for irrigation purposes, the fitness Check with the |
| 246 | European Parliament, regulation on minimum requirements for water reuse (European |
| 247 | Parliament, 2020) is necessary. Since the regulation encourages the reuse of treated urban |
| 248 | wastewater, small-scale decentralized systems could be legally supported in their |
| 249 | implementation. This could result in an integrated water management approach that is also |
| 250 | applicable in rural areas by increasing the sustainability of agricultural irrigation and |
| 251 | providing a reliable alternative to freshwater supply. |
| | |

It should be noted that for small-scale systems it is crucial to adopt the risk-based approach 252 253 outlined in the Water Reuse Risk Management Plans (WRRMPs), which are also included in the proposal 2018/0169 (COD), currently under evaluation by the European boards. When 254 considering small-scale collection and treatment systems, human, technical, and financial 255 256 resources are often limited (WHO, 2006) and thus monitoring strategies of water resources might be challenging. However, the hazard prioritization and risk ranking introduced by the 257 WRRPM could represent a valuable control strategy for a small-scale water system. 258 259 Specifically, control measures can be implemented for the minimum monitoring of community supplies, by monitoring the essential parameters of water quality and thus 260 261 reducing the overall monitoring costs. Furthermore, a legal instrument that can be used as a 262 reference for technical, economic, and environmental aspects is represented in the ISO/TC 282: "Guidelines for Treated Wastewater use for Irrigation Projects" for decentralized 263 systems. Applying the water safety plan to water reuse was examined elsewhere (Goodwin et 264 al., 2015), indicating that similar to the WHO's Framework for Safe Drinking Water, the risk 265 management framework for reuse would guide scheme managers in setting targets and 266 assessing management performance. 267 Most of the published literature on the reclaimed water reuse has focused on the technologies 268

269 and implementations (Capodaglio, 2020; Lee et al., 2018; Rizzo et al., 2020; Salgot and 270 Folch, 2018). In fact, we need more critical analysis and opinion papers on the legislative perspective, such as Rizzo et al. (2018) presented the opinion of the Scientific Committee on 271 272 Health, Environmental and Emerging Risks (SCHEER) on the draft version of the European Commission's "Proposed EU minimum quality requirements for water reuse in agricultural 273 274 irrigation and aquifer recharge" (draft V.3.3, February 2017). The authors suggested that common criteria should be defined for the development of case-by-case assessments, in order 275 to ensure comparable minimum quality requirements across the EU member countries. 276

277 Similarly, principal barriers limiting the reclaimed water use for agriculture in Italy

(particularly in Sicily) were analyzed by Ventura et al. (2019), highlighting the complex and
strict Italian legislations on the reuse of treated wastewater, and commenting that potential
users should rely on the support of private or public agencies such as the Italian Irrigation
Consortia.

Regarding the systems aiming to produce drinking water, the "Drinking Water Directive" 282 (DWD) Council Directive 98/83/EC (and its revision EU 2015/1787) is the starting-point 283 284 legislation for setting actions at the national level. Despite the binding character of the Directive, measures are mandatory to distribution systems serving more than 50 people or that 285 286 provide more than 10 m³ of water per day, while for small-scale systems some exemptions 287 can be applied (Art.3 (2)). Therefore, under the DWD, household and small-scale supply systems (e.g. wells or local springs) for rural communities are not regulated as well as the 288 possibility to produce drinking water from alternative sources (e.g. rainwater and/or water 289 290 vapor). These aspects highlight how the existing EU regulatory instruments for drinking water are not in line with the latest scientific knowledge of the WHO recommendations. 291 Specifically, the Joint Monitoring Program (JMP) by WHO defines rainwater as an 292 293 "improved" water source for potable uses in rural and urban areas, concerning the protection 294 from fecal matter contamination. Moreover, guidelines apply both to large and small-scale 295 piped and non-piped drinking water systems in rural communities and individual dwellings. It should be noted that, even for small-scale collection and treatment systems, no exemptions 296 297 are allowed when potential risks to human health are evident. In this perspective, WHO guidelines provide scientific support, by highlighting the need for a water safety plan (WSP) 298 299 risk-approach for public health protection when small-scale decentralized systems are applied (World Health Organization, 2017). These guidelines (evaluated also in the Proposal EC 300 301 COM 753/2017 (01.02.2018)) provide prescriptions for the safe management, operation, and

monitoring of wastewater, excreta, and greywater in agriculture and drinking-water quality. In 302 most cases, the successful implementation of WSPs is limited by a number of factors such as 303 the lack of financial resources and the absence of legislation (Tsoukalas and Tsitsifli, 2018). It 304 should be noted that when considering small-scale solutions, economic sustainability is a 305 crucial factor. In addition, alternative water sources for drinking water production should also 306 be considered. In this regard, it is fundamental to introduce the main elements of the WHO 307 guidelines in the revision of DWD to provide an EU regulatory instruments which could be 308 309 fully implemented for both centralized and decentralized systems. This means mainly to update the existing safety standards, introduce a risk-based safety assessment, and to include 310 311 measures for drinking water production from alternative sources. Since compliance with 312 water directive might not be sustainable from an economic point of view, enabling the environment for decentralized systems needs to be analysed at national and local levels 313 considering a risk-based approach to water safety. In terms of aspects related to the water 314 supply from unconventional water sources, e.g. recycled and desalinated water, public 315 acceptance is also stated as one of the major barriers in Europe and all over the world(Adapa 316 317 et al., 2016; Hurlimann and Dolnicar, 2016). At the global scale, similarly, there is not a unified regulation for water reuse. Although 318 319 policies on alternative water sources differ between states in the United States, the U.S. 320 Environmental Protection Agency (EPA) introduced the National Water Reuse Action Plan: Collaborative Implementation on February 27, 2020, to develop serious actions on water 321 322 recycling. In Canada, guidelines are released both by the federal government and provincial

323 governments (Van Rossum, 2020). There is only one guideline prepared by the federal

324 government for water reuse in 2010, named as Canadian Guidelines for Domestic reuse Water

325 for Use in Toilet and Urinal Flushing. Whereas Alberta Provinces released a fact sheet on

326 Alternative Solutions Guide for Small System Water Reuse, Atlantic Provinces have a

Wastewater Guidelines Manual including informations for reuse applications. British
Columbia is the only province with the regulation for water reuse. Moreover, in Australia,
water guidelines were published in two phases as a part of the National Water Quality
Management Strategy. The first phase includes a framework and guideline for managing
health and environmental risks including recycled water quality and guidance on the use of
treated sewage and greywater, the second phase focuses on the augmentation of drinking
water supplies, aquifer recharge, and stormwater harvesting and reuse (NRMMC, 2006).

334

3.2. Sludge treatment and reuse for food safety

Organic matter and nutrients are the two main elements that make the use of treated sludge 335 336 suitable for soil fertilization (European Commission, 2019e). Also, the 86/278/EEC "Sewage 337 Sludge Directive" (SSD) (and its in-force revision Regulation (EC) No 219/2009) promotes the application of treated sewage sludge in agriculture (Art.3(2)) as long as the Member States 338 implement necessary measures for protecting human and environmental health and preventing 339 harmful effects on soils (Art.6(a), Art.7). The treatment/recovery of sludge and its reuse 340 through agricultural applications is indeed a major barrier in Europe. In fact, when looking at 341 national legislation, each country has different thresholds, and within certain countries, even 342 343 individual states/provinces may have different threshold values.

344 When considering applications of small-scale treatment systems to decentralized contexts, rural community or neighborhood-based solutions should be implemented for treating the 345 sludge. In these cases, fecal sludge could represent the relevant input for producing compost. 346 347 In this regard, the SSD sets limits on the concentrations allowed (in soil and sludge) for the application of the residual sludges from septic tanks (Art.2 (a -i, -ii, -iii)). Consequently, no 348 349 explicit barriers are detected for the replicability of small decentralized systems. Despite this, the European Legislative Framework lacks ad-hoc regulation for community-based 350 composting and co-composting systems to fully support and regulate the recovery and reuse 351

of sludge in small and rural communities. Furthermore, the SSD often receives criticism as
being outdated and does not include limits for pathogens and organic micropollutants in soil
and sludge.

Considering the compost production, a further aspect is to analyse the possibility of labeling 355 and marketing of the fertilizer as an EU product. In this perspective, the European Fertilizer 356 Regulation (2009b, No 1069/2009) limits the exploitation of small decentralized systems, 357 since compost derived from digestate and sewage sludge cannot be labeled and marked as EU 358 359 fertilising products (Annex II). A possible way to address this barrier is currently provided by different valuable works and projects. In 2019, the report "Digestate and compost as 360 361 fertilizers: Risk assessment and risk management options" was published by Wood with 362 partners Peter Fisk Associates and Ramboll for European Commission (Corden et al., 2019). In this work, no limitation on input materials or uses is detected for compost or digestate 363 when an environmental and human health risk assessment and a risk management options 364 analysis (RMOA) are implemented. In a JRC Science for Policy report (JRC, 2019) possible 365 legal framework for marketing fertilising products, derived from precipitated phosphate salts, 366 thermal oxidation materials and pyrolysis, gasification materials, and derivates (STRUBIAS), 367 is explored. The main attention in STRUBIAS material is given to phosphate salts, which can 368 369 be obtained from wastewater and sewage sludge from municipal WWTPs by AD or by composting. This study shows how STRUBIAS can be considered as a valuable framework to 370 provide phosphorus in a safe way to reduce the demand for the primary raw material from 371 372 phosphate rocks. Hence, the compost produced within decentralized systems might decrease the demand for synthetic fertilizers and reduce economic/environmental impacts associated 373 374 with fertilizer production and waste disposal also in rural areas. In Europe, the European Sustainable Phosphorus Platform (ESPP) serves as a hub for information exchange and 375 376 facilitates communication between all cross-sectoral stakeholders. In fact, political interest in

377 phosphate sustainability has grown a lot at the European level. Incorporation in the EU 378 critical materials list is seen as vital in this respect (de Boer et al., 2018). European legislation governing phosphorus recycling was critically reviewed by Hukari et al. (2016), where 379 legislation harmonisation, the inclusion of recycled phosphorus in existing fertiliser 380 regulations, and support of new operators were proposed to speed up market penetration of 381 novel technologies, reduce phosphorus losses and safeguard European quality standards. 382 When using compost produced from waste and/or irrigate the site with reclaimed water for 383 384 food crops, food safety is a crucial factor to assess. Namely, the fitness check with Commission Regulation (EC) No. 1881/2006 on maximum levels for certain contaminants in 385 386 foodstuffs should be analysed. According to this regulation, no relevant barriers were detected 387 for the marketability of products irrigated with reclaimed water and/or fertilized with compost from waste, as the compliance depends exclusively on the final product. Major constraints 388 were detected when considering organic farming. In fact, according to the Organic Farming 389 390 Regulation (EC) No. 889/2008 (and its revision Regulation (EU) No. 848/2018) no information is provided regarding sewage sludge matrix for fertilizer production. Thus, the 391 sewage sludge cannot be used to improve soil quality. However, "composted or fermented 392 393 household waste" can be authorized as long as it contains only vegetable and/or animal waste. 394 For instance, in the study of Viaene et al. (2016), among the 28 identified barriers to on-farm 395 composting and compost application, the complex regulation was listed as one of the main five barriers. In fact, the authors recommended a certain degree of flexibility in current 396 397 policies and institutional arrangements to stimulate compost production and application.

398

3.3. Renewable resources exploitation for energy efficiency

To close the loop of the Integrated Resources Management, the energy sector should also be assessed. In this regard, small-scale decentralized systems can offer a valuable alternative to methane extraction from natural deposits, such as upgrading methane from biogas produced

402 in anaerobic treatment. With the view to reuse methane in the transport sector, small-scale decentralized systems are supported by the renewable energy directive 2018/2001/EU (EC, 403 2018) to meet the 10% of renewable resources used in transport. Specifically, in Annex IX(f) 404 sewage sludge can be used to produce biogas for transport and advanced biofuels. It should be 405 406 considered that biogas and biomethane should ensure the quality requirements as defined in the technical standards for biogas EN 16726 and EN 16723 on Liquefied Natural Gas (LNG), 407 biomethane, and blends for automotive fuels. Concerning the investigated barriers for the 408 implementation of technologies promoting biogas reuse, many studies highlighted the lack of 409 institutional support and specific action programs to support biogas technologies (Nevzorova 410 411 and Kutcherov, 2019). In this regard, complex institutional and legal pathways could block 412 and prevent the implementation of such applications. Furthemore, Yaqoot et al. (2016) analysed barriers to the dissemination of decentralized renewable energy systems. Among the 413 institutional barriers, the authors include the lack of a suitable legal and regulatory framework 414 for dissemination of decentralized renewable energy systems as a major institutional barrier, 415 together with the other sub-barriers such as the lack of agencies to disseminate information, 416 417 uncertain government policies, strict bureaucratic procedures, unstable macro-economic 418 environment, lack of stakeholder participation in decision making, clash of interests among 419 stakeholders; lack of R&D culture; insufficient professional institutions and lack of private 420 sector participation.

421 **4. Regulatory fitness check**

To check and outline available conditions or possible obstacles in the implementation of the small-scale decentralized collection and treatment systems within the European legislative framework, Evaluation Fitness Check is reported in **Table 2**. The main parameters of the solutions, intended as key factors to get the Legislative compliance, were highlighted concerning the prescriptions of the policies analysed (see column "parameters to consider").

- 427 For each parameter, the reference documents were listed (see column "reference documents").
- 428 Documents were grouped according to directives, technical standards, guidelines, manual (see
- 429 column "document type"). Finally, "relevant information" was reported to point out whether
- the regulatory instruments "support" or "hinder" the recovery and use/reuse of specific
- 431 resources/by-products.
- 432

433 Table 2. Summary of fitness check for small-scale decentralized systems (A: "considered no barrier", in green: if quality and/or safety standards

434 are met and the output reuse is generally allowed; B: "not considered", in yellow: no explicit reference/information in the legislation; C:

435 "considered, potential barrier" in red: if legislation highlights possible constraints to be overcome).

| Application field | Parameters to consider — | Reference documents | | Relevant |
|----------------------------------------|-----------------------------------------------------------------------------|---------------------|--------------------------------------------------------------------------------------------------|-------------|
| Application field | | Document type | Reference number | information |
| | Categories of crops allowed to be cultivated | | Proposal 337/2018 and 2019 revision ISO/TC 282 | А |
| From wastewater to reclaimed water for | Indicative treatments required for reaching water quality effluent | Directive _ | Proposal 337/2018 and 2019 revision ISO/TC 282 | А |
| irrigation | Influence of flow quality on the final intended use | | Proposal 337/2018 and 91/271/EEC and revisions ISO/TC 282 | А |
| | CE Labelling for the fertilizer | | 1009/2019 | С |
| | Influence of sewage sludge for agricultural uses | Directive | 219/2009 | А |
| | Use of sewage sludge for organic farming | | 889/2008 | С |
| From sewage sludge to compost | Compost parameters to be considered for agricultural uses at National Level | Report | ENV.A.2. /ETU/2001/0024 (Amlinger et al., 2004) Digestate and compost report JRC Report | А |
| | Use of sewage sludge for agricultural purposes | | ENV.A.2. /ETU/2001/0024 Digestate and compost report JRC Report | А |
| From biogas to biomethane as | Biomethane characteristics for transport, distribution and use | Technical standards | EN 16726 | А |
| automotive fuel | Characteristics of methane for use as automotive fuel | | EN 16723-2 | А |
| Agroforestry system | Categories of crops allowed | Directive | Proposal 337/2018 and 2019 revision | А |

| | Indicative treatments required for reaching water quality standard | | Proposal 337/2018 and 2019 revision | А |
|----------------------------------------|----------------------------------------------------------------------------|--------------------------------|---------------------------------------------------|---|
| | influence of flow quality on the final intended use | | Proposal 337/2018 and 91/271/EEC and revisions | А |
| | Types of food regulated | | 1881/2006 | А |
| | Limits for compliance in foodstuffs | | 1881/2006 | А |
| | Amount of water required by crops | Manual | FAO Manual | В |
| | Permitted reuse of rainwater in relation to required treatments | Guideline | Guidelines and report from world-wide experiences | А |
| | Permitted water source for drinking purpose | | 2000/60/CE and 98/83/EC and revisions | В |
| From rainwater to | Parameters to meet for potable uses | Directive | 98/83/EC and revision; (Proposal 753/2017 EC) | В |
| drinking water | Permitted water source for drinking purpose | WIIO Cuidalings for drinking y | WHO Guidelines for drinking water | А |
| | Treatment for the purpose and monitoring/control measure | Guidelines | quality | В |
| | Type of water to aquifer recharge | | 2000/60/CE | А |
| From rainwater and | Attention to water quality to maintain the "Good" status of groundwater | | 2000/60/CE and 2006/118/EC (2014/80/EU) | А |
| runoff from road to | Categories of crops allowed to be cultivated | Directive | Proposal 337/2018 and 2019 revision | А |
| water for aquifer recharge and further | Indicative treatments required for reaching water quality effluent | | Proposal 337/2018 and 2019 revision | А |
| irrigation | influence of water quality for irrigation use | | Proposal 337/2018 and 2019 revision | А |
| | Permitted reuse of rainwater concerning required treatments | Guidelines | Guidelines and report from world-wide experiences | А |
| From brine to salt production | Salt quality parameters | Standards | CXSTAN 150-1985 (Codex Standard, 1985) | В |
| | Water source for drinking purpose | Directive | 2000/60/CE and 98/83/EC and revisions | В |
| From water vapor to drinking water | Parameters to respect for potable uses | | 98/83/EC and revision; Proposal 753/2017 EC | В |
| urmking water | Treatment for the purpose and monitoring/control measure | Guidelines | WHO Guidelines | В |

When considering irrigation for agroforestry, crop categories, indicative required treatments, 439 and influent water quality need to be considered within the EU directives and the ISO/TC 440 282. The application necessities should be evaluated within the EC 337/2018, EC 271/1991, 441 and EC 118/2006. Besides, the typical amount of water should be guaranteed according to the 442 type of crop, as specified in the FAO Manual. When considering compost spreading, on the 443 other hand, no unified EU regulation/legislation/directive is specified. However, quality 444 parameters and/or presence of sewage sludge were evaluated in several Reports (Amlinger et 445 al., 2004; Corden et al., 2019; JRC, 2019). Concerning biomethane production, quality 446 parameters for automotive fuel applications are specified in EU level directives. In terms of 447 448 rainwater and/or runoff treatment and reuse for irrigation, since no specific prescriptions are 449 defined in the EU legislative framework, further analysis is required at the national/local level. Concerning the drinking water, quality standards cannot be applied to small-scale water 450 451 supply and sanitation systems. When aquifer recharge is involved, the discharged water is regulated by the WFD in terms of quality to maintain a "good" status of groundwater. 452 Regarding salt production, quality standards defined by FAO and WHO needs to be ensured 453 when food-grade salt is considered. Since no information on the source of salt and the 454 455 minimum treatment is provided, reuse is outlined as "not defined", while national/regional 456 regulations should be further analysed.

457 **5. Financial analysis**

The choice for funding water infrastructure between a centralized and decentralized solution
in a rural and/or peri-urban area depends on several variables: method of economic
assessment (social discount rate), funding policy (funding rate), and users' self-organization
(cost-sharing) (see the paper, (Brunner and Starkl, 2012)). In fact, a successful
implementation of decentralized solutions relies on many critical factors such as public
acceptance, qualified maintenance, organizational support, and availability of financial

resources (Sousa-Zomer and Cauchick Miguel, 2018). The nature of financing arrangements 464 that depends on the institutional structure was assessed to define a general structure of the 465 mechanisms on which the water/wastewater management is based. The approach proposed by 466 the Organization for Economic Cooperation and Development (OECD) focuses on the 3T 467 (Taxes, Tariffs, Transfers) for regulating, increasing, and balancing finances in three forms. 468 The structure of financing pathways for small-scale decentralized technologies was analyzed 469 by highlighting the general framework for ensuring cost recovery (Fig. 1) of municipal water 470 cycle services (e.g. wastewater and domestic water). 471

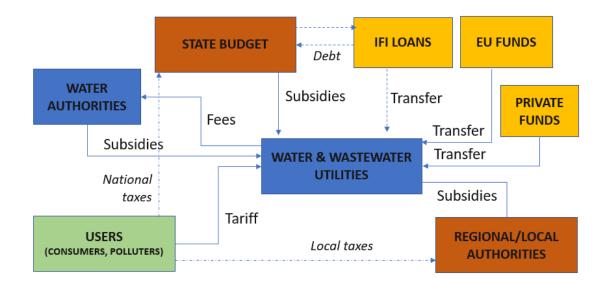




Fig. 1. General structure of the financial framework in water service management.
The analysis conducted for all EU member states highlighted that the cost recovery is mostly
achieved through tariffs, whose affordability differs from between countries. Specifically, the
majority of the Countries use the taxpayer's money to cover the Capital Expenditure costs
(CAPEX), while few economically weaker Countries use foreign funds as "money transfers"
for cost recovery. Moreover, the source of used subsidies for cost recovery can be divided
into two main categories (Table 3):

Countries with a specific water financing structure. In this case, the concept of "water
pays water" is followed;

- Countries with financing strategies mainly derived from public budget at national or
- 483 local/regional levels.

| Countries | National administration | Water authority | Local / Regional authorities |
|-----------|-------------------------|-----------------|------------------------------|
| Croatia | Х | X | X |
| Spain | Х | X | X |
| Cyprus | Х | Х | |
| France | Х | Х | |
| Germany | | Х | |
| Poland | Х | Х | |
| Belgium | | | Х |
| Austria | Х | | Х |
| Bulgaria | Х | | Х |
| Greece | Х | | Х |
| Italy | Х | | Х |
| Portugal | Х | | |

484 **Table 3.** Summary of subsidies source.

485

It was also highlighted that for small-scale decentralized treatment systems, whose owners
and suppliers could be either a public or private body, a blend of three financing strategies
(e.g. Tariff, Subsidies, Transfers) could be implemented to cover the initial investment and
thus reduce the Return On Investment (ROI) period. The details are shown in **Table 4** below.

490 **Table 4.** Summary of financing sources.

| Service | Type of costs | Tariff | Subsidies (taxpayers' | Transfers |
|----------------|---------------|--------|-----------------------|-----------|
| delivered | | | money) | |
| Wastewater | CAPEX | Х | Х | X |
| treatment | OPEX | Х | Х | |
| Domestic water | CAPEX | Х | Х | X |
| | OPEX | Х | Х | |
| Agriculture | CAPEX | Х | Х | X |
| water | OPEX | Х | | |

In all these cases, in which the delivered service (e.g. water supply or wastewater treatment) 491 does not involve any financial transaction, end users will be the payers of the "tariff" (e.g. a 492 farmer will pay to maintain/operate its small-scale decentralized treatment system and the 493 494 initial investment at least partly).

Furthermore, local, regional or national public bodies or water service operators can provide 495 subsidies especially when costs for drinking water supply (e.g. tap water) and sanitation 496 services through public networks are unaffordable. These measures can be applied to provide 497 498 universal and equitable access to drinking water and sanitation to all. Transfers (in the form of foreign or EU funds) are mainly used to cover investment costs for building or revamping a 499 500 water infrastructure, targeting vulnerable and less developed areas in the EU and worldwide.

501 6. Possible solutions: a step forward in the context of the Innovation Deal and the 502

European green deal

The aspects analysed in this study can be contextualized in a broader framework such as the 503 European Green Deal (European Commission, 2020). Achieving a zero-climate impact by 504 2050 will only be possible if measures to be adopted are supported by adequate European 505 regulatory instruments that can be applied both at large- and small-scale. In this context, the 506 507 European regulatory framework highlighted signs of disparity. Clear policy instruments 508 regarding sustainable solutions in small and rural communities or agglomerations are 509 currently lacking. The result of this gap is the achievement of quality standards that could represent a challenge for the economic sustainability of decentralized systems. 510 511 Moreover, according to the last revision of the UWWTD (91/271/EC), pollution from urban wastewater systems to water and soil can still be avoided. Specifically, sources of pollution 512 513 are related on one side to unmonitored/untreated combined sewer overflows, small

agglomerations, and non-connected dwellings and on the other side to possible toxic and 514

emerging contaminants in sewage sludge used in farming. 515

The challenge is to consider the health and environmental risk related to the emerging 516 517 contaminants (e.g. pharmaceuticals, micro-nanoplastics) while recovering and safely reusing water and raw materials. Therefore, the implementation of small-scale treatment solutions must 518 be supported by ad-hoc regulations for small agglomerations and by a regional integrated health 519 520 and environmental risk-based approach to sustainably manage solid and liquid non-zeropolluted residual streams. In this regard, innovation actions should be promoted, such as the 521 522 Innovation Deal (Innovation Deal, 2017), to support European and national governments in proposing and adopting policies more oriented towards rural services. Recently, Jiménez-523 Benítez et al. (2020) assessed the reclaimed water reuse in fertigation via AnMBR technology 524 525 within the EU Innovation Deal. The authors highlighted that in order to take full advantage of 526 the benefits of AnMBR technology on water reuse, favorable and harmonized regulations among the EU States would need to be adopted. 527

Similar to the approach of the community composting regulation, Innovation Principle 528 (Innovation Principle, 2017) can be implemented to design a policy framework that promotes 529 the sustainable management of water and water-related services also in rural areas at a 530 community-based level. For instance, similarly to the Italian community composting draft 531 proposal, small-scale technologies could be formally implemented either by the municipality 532 533 in co-creation with citizens or by a "collective body" (e.g. two or more domestic/non-534 domestic users established in condominium, association, consortium, etc.). Specifically, in the latter case, the implementation should start after the collective body sends a certified 535 536 notification of the activity to the competent municipality, which in turn notifies the collective body with the management service. 537

538 When considering the management of small-scale systems, the activity could be carried out

either by the municipality or by the collective body. Regardless of the manager, data

540 regarding the system treatment efficiency, amount of produced wastes, effluent characteristic

- for quality monitoring should be registered. Monitoring strategies should be anyway carried
 out and supervised by the competent authority (e.g. province, municipality, etc.).
- 543 Thus, the community composting based-approach could be implemented in the water and
- 544 water-related sector to support the implementation of small-scale treatment technologies as
- 545 according to the scheme represented in **Fig. 2**.

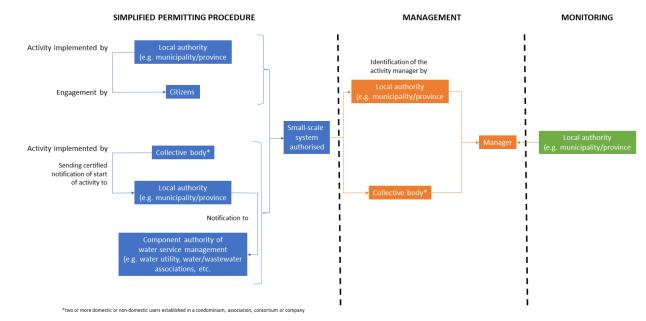


Fig. 2. Implementation of Community Regulation in the context of small-scale collection and
treatment systems.

549 Furthermore, the regulatory gap that must be bridged to fully support the integrated

550 management of resources should include measures aimed at providing quality standards and

- 551 monitoring procedures suitable for small-scale collection and treatment systems as they are
- more vulnerable to breakdown and contamination than larger utilities. These systems are also
- the most sensitive in terms of economic sustainability. Therefore, the application of
- regulations for centralized systems may not be economically sustainable, thus representing a
- 555 constraint or a barrier for further development.
- 556
- 557

558 **7.** Conclusion

Although no general barriers are detected for the reuse of reclaimed water in the EU 559 legislation, there are some limitations on drinking water production from non-conventional 560 water resources. On the other hand, major constraints are determined for EC-marked compost 561 for organic farming. Since no relevant information is found for the community composting at 562 the EU level, regulatory instruments for policy support should be analysed at the local level. 563 As a result, the current EU legislative framework does not provide ad-hoc guidelines to close 564 the water loops for a small decentralized system, highlighting a lack of the enabling 565 environment for small-scale decentralized technologies at the community level. Moreover, 566 567 some gaps are determined in terms of regulatory framework, institutional support, financing 568 schemes for small and rural communities, which might hinder the implementation of decentralized systems. In this regard, possible blockages for the exploitation of small-scale 569 570 treatment solutions could be the achievement of the quality standards, as set out in EU Directives, which might not be economically feasible for these systems. To deal with the 571 European Green Deal challenge, sustainable growth in terms of social, economic, and 572 environmental progress should be delivered by adopting Innovation Principle. Concerning the 573 574 financial aspects, water tariff structures are mainly addressing urban environment and larger 575 utilities' needs, and smaller service authorities have to find ad-hoc solutions for local service 576 providers. To develop a full picture of water and water-related small and decentralized services to deliver regenerated closed loops, an innovation deal might be prepared to support 577 578 European (and national) governments.

579 Credit authorship contribution statement

580 Giulia Cipolletta: Investigation, Methodology, Formal Analysis, Visualization, Writing -

581 original draft. E. Gozde Ozbayram: Investigation, Writing - original draft. Anna Laura

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- 584 Writing review & editing. Eric Mino: Conceptualization, Visualization, Writing review &
- 585 editing. Francesco Fatone: Funding acquisition, Project administration, Resources,
- 586 Supervision, Writing review & editing.

587 **Declaration of competing interest**

- 588 The authors declare that they have no known competing financial interests or personal
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- 596 Supplementary data to this article can be found online

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