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Toxic marine microalgae and noxious blooms in the Mediterranean Sea: A contribution to the Global HAB Status Report

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25 **ABSTRACT**

26 We review the spatial distribution of toxic marine microalgal species and the impacts
27 of all types of harmful algal events (Harmful Algal Blooms, HABs) in the
28 Mediterranean Sea (MS), including the Black Sea, the Sea of Marmara, coastal
29 lagoons and transitional waters, based on two databases compiled in the Ocean
30 Biogeographic Information System (OBIS). Eighty-four potentially toxic species have
31 been detected in the MS (2,350 records), of which 16 described from these waters
32 between 1860 and 2014 and a few suspected to have been introduced. More than half
33 of these species (46) produce toxins that may affect human health, the remainders
34 ichthyotoxic substances (28) or other types of toxins (10). Nevertheless, toxicity-
35 related events are not frequent in the MS (308 records in 31 years), and mainly consist
36 of impacts on aquaculture, caused by the dinoflagellates *Dinophysis* and *Alexandrium*,
37 along with a few actual shellfish poisoning cases. *Pseudo-nitzschia* blooms are
38 widespread, but domoic acid in shellfish rarely exceeds regulatory levels. Fish kills
39 are probably less sporadic than reported, representing a problem at a few places along
40 the southern MS coasts and in the Ebro River Delta. Since the last decade of the 20th
41 century, blooms of the benthic dinoflagellates *Ostreopsis cf. ovata* have regularly
42 occurred all along rocky shores of the MS, at times with human health problems
43 caused by toxic aerosol. New records of *Gambierdiscus* and *Fukuyoa*, until now
44 reported for the westernmost and easternmost MS coasts, raise concerns about the risk
45 of ciguatera, a syndrome so far known only for subtropical and tropical areas. Recent
46 discoveries are the dinoflagellates *Vulcanodinium rugosum*, responsible for the
47 presence of pinnatoxins in French lagoons' shellfish, and the azaspiracid-producers
48 *Azadinium* spp. Mucilages and discolorations have a major impact on tourism in
49 summer. Reports of toxic species and HABs have apparently increased in the MS over

50 the last half century, which is likely related to the increased awareness and monitoring
51 operations rather than to an actual increase of these phenomena. Indeed, while the
52 case of *Ostreopsis* appears as a sudden upsurge rather than a trend, no actual increase
53 of toxic or noxious events has so far emerged in intensively studied areas, such as the
54 French and Spanish coasts or the Adriatic Sea. Moreover, some cases of decrease are
55 reported, e.g., for *Alexandrium minutum* blooms disappearing from the Harbour of
56 Alexandria. Overall, main HAB risks derive from cases of massive development of
57 microalgal biomass and consequent impacts of reduced coastal water quality on
58 tourism, which represents the largest part of the marine economy along the MS coasts.

59

60 **Keywords:** HABs; Mediterranean Sea; microalgae; toxicity; OBIS

61

62 **1. Introduction**

63 The Mediterranean Sea (MS, from the Latin *mare Mediterraneum* = the sea
64 surrounded by land) is an enclosed basin surrounded on the north by southern Europe
65 and Anatolia, on the south by North Africa and on the east by the Levant. It occupies
66 an area of approximately 2,510,000 km² lying between latitudes 30° and 46° N. The
67 narrow and shallow Strait of Gibraltar to the West connects it with the Atlantic
68 Ocean, the Dardanelles to the East with the Black Sea through the Sea of Marmara
69 and the Bosphorus, while to the south-east the Suez Canal, opened in 1869 and recently
70 expanded, allows the exchange with the Red Sea. In spite of its geographic position
71 within the northern temperate latitudes, the quite shallow sill (170 m) at the Atlantic
72 boundary blocks the entrance of deep, cold oceanic waters and determines temperate-
73 subtropical conditions in the whole area, with minimum temperatures rarely and only
74 at certain locations going below 12 °C.

75 The size, location, and morphology of the MS are at the base of its complex physical
76 dynamics with a distinctive thermohaline circulation and permanent or semi-
77 permanent sub-basin gyres. A marked oligotrophy, increasing along both the west-
78 east and the north-south directions, characterizes the MS (Siokou-Frangou et al.,
79 2010). However, along the Mediterranean coasts there are densely populated areas
80 while a number of large rivers with extended catchment basins flow in the MS (e.g.,
81 the Po in the northern Adriatic, the Nile in Egypt, the Ebro in Spain, and the Rhone in
82 France). This implies that meso- and eutrophic conditions, and at times pollution, can
83 affect various coastal areas (UNEP/MAP, 2012).

84 The MS has been the crossroad of various cultures since the very beginning of the
85 human colonization and the development of ancient civilizations. Trading routes,
86 migrations, invasions and the struggle for power have shaped the dynamic history of

87 populations around the basin for millennia. The population grew from 281 million in
88 1970 to 419 million in 2000 and 472 million in 2010, and is predicted to reach 572
89 million by 2030. Coastal administrative entities make less than 12% of the surface
90 area of the Mediterranean countries, but host more than a third of the population of
91 the whole region. Coastal population grew from about 100 million in 1980 to 150
92 million in 2005 and could reach 200 million by 2030 (UNEP/MAP, 2017).

93 The MS also represents a unique geographic landscape that generates wealth but
94 requires cooperation among the different countries to preserve the environment and
95 the biological resources. The conservative value of the economic assets of the MS has
96 been estimated to be in the order of US\$ 5.6 trillion, generating an annual economic
97 value of US\$ 450 billion (Randone et al., 2017). A large fraction of the economic
98 value is represented by tourism and related activities; fisheries come as second but
99 >80% of the fish stock is presently threatened. Aquaculture in the MS has
100 considerably expanded over the last decades reaching about 1.3 million tons in 2009
101 with an estimated value of US\$ 3,700 million (Rosa et al., 2012). Most of the marine
102 aquaculture production comes from the north Mediterranean countries, which are also
103 the most intensively monitored, but it is rapidly expanding also in Turkey and Egypt.

104 In spite of the dramatic alteration of habitats, depletion of natural resources and
105 increased number of alien species, the MS is still characterized by high biodiversity in
106 most animal and algal groups and a considerable number of endemic species (Coll et
107 al., 2010).

108 The rate at which climatic conditions (e.g., surface temperature, heat waves and sea
109 level) have changed in the MS over the last decades is higher than the global average
110 (Cramer et al., 2018). These changes, coupled with increased population size,
111 urbanization and changes in land use at many coastal places, may pose at serious risk

112 the quality of the environment, the quality and quantity of food and consequently the
113 health and safety of the local populations (Cramer et al., 2018). Especially in view of
114 the growing need to exploit marine resources, HABs and toxic species may represent
115 an increasing risk for human health and economic activities.

116 Few are the papers reviewing the occurrence of harmful species and/or events at the
117 scale of the whole Mediterranean basin. Fifty years ago Jacques and Sournia (1978-
118 1979) published a first account of the cases of water discoloration ('eaux rouges') and
119 the species involved. The overview included mainly dinoflagellate blooms, along with
120 a few cases of anoxia but with no evidence of toxic effects in humans or marine fauna
121 in those years when microalgal toxins were still almost unknown. In an overview of
122 nearly twenty years later, cases of PSP and DSP – mainly attributable to *Alexandrium*
123 *minutum* and *Dinophysis* spp., respectively – were reported from the northern coasts
124 of the basin, along with the records of various potentially toxic or ichthyotoxic
125 dinoflagellates at different sites (Honsell et al., 1995). A subsequent overview of toxic
126 and harmful microalgae covering up to 2009 pointed at the sudden spreading of
127 *Ostreopsis* cf. *ovata* blooms along the rocky Mediterranean shores (Zingone, 2010).

128 The present overview covers the MS distribution of marine, toxin-producing
129 microalgae, as included in the IOC-UNESCO Taxonomic Reference List of Harmful
130 Micro Algae (Moestrup et al., 2009) and the cases of toxin-related harmful events
131 (Sections 2.1 and 2.2), including direct impact on human health or natural resources
132 or indirect impact to aquaculture industry. In addition, we review non-toxic events
133 that include high biomass harmful algal blooms (HB-HABs) causing seawater
134 discolorations, anoxia or any other damages to the environment or human activities
135 (Section 2.3). Finally, we discuss the trends of HABs in the MS in general and
136 particularly in the Adriatic Sea, which is considered a HAB hotspot (Section 3). The

137 overview is based on information from more than 600 scientific publications and
138 technical reports collected in two curated databases in the Ocean Biogeographic
139 Information System OBIS (Zingone et al., submitted): the MS-HABMAP-OBIS
140 (<https://obis.org/>), gathering records of toxic species occurrence, and the Harmful
141 Events Database (HAEDAT, <http://haedat.iode.org/>), collecting information of either
142 toxic or non-toxic events, i.e., cases of intoxications, closures of aquaculture plants,
143 seawater discolorations and mucilages. The present review is a contribution to a first
144 appraisal of the current knowledge of HAB occurrences across the world seas,
145 namely, the Global HAB Status Report, (Hallegraeff et al., 2017; Zingone et al.,
146 2017). The requirement for such an assessment has emerged from the apparent
147 worldwide increase and spreading of HABs and their negative impacts contrasted by
148 the lack of an overview founded on a robust basis of data.

149

150 **2. HABs in the Mediterranean Sea: toxic species and harmful event distribution**

151 *2.1 Toxic species*

152 Of the more than 140 potentially toxic species listed in the IOC-UNESCO taxonomic
153 reference list (Moestrup et al., 2009), 84 have been found in the MS so far: 17
154 diatoms, 54 dinoflagellates, 3 dictyochophytes, 6 haptophytes, and 4 raphidophytes
155 (Table 1, and some examples in Fig. 1). These records cover both species actually
156 found to produce toxins in the MS and species known to be toxic from other areas.
157 Given the known variability in toxin production among strains of the same species,
158 non-tested local populations are only ‘potentially toxic’ in most cases, but for brevity
159 they will be referred to as ‘toxic’ in the context of this paper. Sixteen of the toxic
160 species have actually been discovered and described from the MS (Table 2), the first
161 ones (*Prorocentrum lima*, *Dinophysis caudata*, *D. sacculus* and *D. tripos*) in the

162 second half of the 19th century and the most recent ones (*Vulcanodinium rugosum*,
163 *Azadinium dexteroporum*, *Nitzschia bizertensis* and *Ostreopsis fattorussoi*) in the
164 current decade. Some of the HAB species of the MS, such as *D. caudata* and
165 *Chattonella subsalsa*, are widely distributed worldwide while others, including the
166 recently described *N. bizertensis* and *O. fattorussoi*, so far seem to be restricted to
167 specific areas of the MS.

168 The discovery of potentially toxic species in the MS has undergone an evident
169 escalation over the years (Fig. 2), from the first descriptions of more than a century
170 before the discovery of their toxicity to the rapid increase after the 1960s and the most
171 recent findings. Information on their distribution has also markedly increased along
172 with the intensification of monitoring operations and studies on planktonic and
173 benthic microalgae (e.g., Zingone et al., 2006; Aligizaki et al., 2009; Pistocchi et al.,
174 2012; Balkis and Taş, 2016; Fernández et al., 2019) and of their resting stages in the
175 sediments (Bravo et al., 2006; Satta et al., 2013) or sediment traps (Montresor et al.,
176 1998). Yet the actual range of most toxic species in the MS is far from being known.
177 Indeed, the identification of some of the most represented genera in the MS, such as
178 *Alexandrium*, *Karenia*, *Karlodinium* and *Pseudo-nitzschia*, as well as of many other
179 flagellates, is quite problematic. In many cases the observation of live material or
180 methods more complex than light microscopy are needed. Cryptic diversity
181 discovered in many microalgal taxa over the last decades also concerns several
182 harmful genera and species, which have undergone careful taxonomic investigations
183 more than other non-toxic taxa. This trend has led to the discovery of non-toxic taxa
184 morphologically similar to toxic ones, such as several species in the *P. delicatissima*
185 and *P. pseudodelicatissima* species-complexes (Bates et al., 2018), the non-toxic *A.*
186 *tamutum* hardly distinguishable from *A. minutum* (Fig. 1A, Montresor et al., 2004),

187 and the non-toxic chain-forming *Gymnodinium impudicum* (as *Gyrodinium*
188 *impudicum*, Fraga et al., 1991) which was misidentified as *Gymnodinium catenatum*
189 in studies predating its discovery (e.g., Carrada et al., 1991). Recent studies coupling
190 detailed morphological investigations with the analysis of different molecular markers
191 and toxin production have attempted to clarify species identity within the
192 *Alexandrium tamarense*-species complex (John et al., 2014; Litaker et al., 2018). The
193 case of *Chattonella subsalsa* is interesting because, based on several molecular
194 markers, two different genotypes with different geographic distributions exist for the
195 species (Klöpffer et al., 2013). All these taxonomic insights have invalidated many
196 previous identifications of presumed toxic taxa, as detailed in the following sections.
197 In recent years, information on the presence of toxic species is also gathered through
198 molecular identification of environmental DNA samples (e-DNA metabarcoding),
199 which may give relevant information on the presence and seasonality of cryptic or
200 rare species (Ruggiero et al., 2015; Dzhembekova et al., 2017; Grzebyk et al., 2017).
201 Nonetheless, new findings of species through molecular methods should always be
202 confirmed by morphological studies.

203 Some of the toxic species of the MS have been suspected to be non-indigenous
204 species (NIS), i.e., introduced outside their natural past or present distribution. The
205 main possible NIS in the MS are *Pseudo-nitzschia multistriata*, *Alexandrium*
206 *pacificum* and *Ostreopsis* cf. *ovata*. The first MS record of *Pseudo-nitzschia*
207 *multistriata*, a chain-forming diatom having a distinctive sigmoid shape (Fig. 1G),
208 was in 1992 in the Gulf of Naples, where phytoplankton have been intensively studied
209 since the beginning of the 1980s. The species has shown an increasing trend
210 afterwards in the same area (D'Alelio et al., 2010) and has subsequently been found in
211 Spanish (Quijano-Scheggia et al., 2008), Greek (Moschandreu and Nikolaidis, 2010),

212 Tunisian (Sahraoui et al., 2011) and Moroccan waters (Rijat Leblad et al., 2013) and
213 in the Adriatic Sea (Pistocchi et al., 2012; Turk Dermastia et al., 2020). The chain-
214 forming dinoflagellate *Alexandrium pacificum* (as *A. catenella*) was found for the first
215 time in low density in 1983 along the Spanish coast (Margalef and Estrada, 1987). In
216 the following years, *A. pacificum* formed blooms on the Spanish coast (Gomis et al.,
217 1996; Vila et al., 2001) and in the Thau Lagoon (as *A. tamarense/catenella*, Abadie et
218 al., 1999; Lilly et al., 2002). Subsequently it was progressively found eastward along
219 the Italian (Lugliè et al., 2003, 2017; Satta et al., 2013), Algerian (Frehi et al., 2007)
220 and Tunisian coasts (Turki and Balti, 2007; Fertouna-Bellakhal et al., 2015), whereas
221 it is still unrecorded in the rest of the MS. The benthic dinoflagellate *Ostreopsis* cf.
222 *ovata* showed a sudden emergence in the MS at the end of the last century (see section
223 2.2.4). A much higher genetic variability and several cryptic species characterize this
224 taxon along the Japanese coasts compared to the Mediterranean-Atlantic area (Sato et
225 al., 2011; Penna et al., 2012) where genetic differences are seen only at the population
226 level with AFLP markers (Italiano et al., 2014). This situation suggests a relatively
227 recent radiation of the species in the latter area and, given the lack of hydrographic
228 links between the two regions, a possible man-mediated transport, although it is
229 impossible to establish when this occurred (Sato et al., 2011). In lack of type material,
230 or material from the type locality, it has not been established which of the numerous
231 morphologically similar taxa corresponds to *Ostreopsis ovata*. Therefore these taxa
232 should be referred to as *O. cf. ovata* (Penna et al., 2010; Sato et al., 2011). Benthic
233 *Gambierdiscus* and *Fukuyoa* species are also a novelty in the MS, and their
234 distribution, presently limited at the two ends of the basin, hints at a possible recent
235 introduction from both the Atlantic and the Red Sea.

236

237 2.2 Toxic events

238 2.2.1 Diarrhetic Shellfish Poisoning (DSP)

239 DSP toxins in mollusks represent the most frequently reported cases of seafood
240 contamination in the MS. Eight toxic species of the genus *Dinophysis*, plus
241 *Phalacroma rotundatum* (Table 1), have been observed along the Mediterranean
242 coasts (Fig. 3A). *Dinophysis caudata* and *D. sacculus* (Fig. 1C), the most frequently
243 reported species, were both described from the MS more than one century ago (Kent
244 1881; Stein 1883), but risks for human health have first been recognised only in the
245 1980s in the Gulf of Lion (Belin et al., 1995). In the northern Adriatic Sea, DSP
246 toxicity events have occurred on both the western and eastern side, often causing the
247 closure of shellfish farms (Sedmak and Fanuko, 1991; Boni et al., 1992, 1993; Della
248 Loggia et al., 1993; Orhanović et al., 1996; Bernardi Aubry et al., 2000; Francé and
249 Mozetič, 2006; Marasović et al., 2007; Ninčević-Gladan et al., 2008). In the period
250 1989-2018, such closures occurred regularly along the Slovenian coast (northern
251 Adriatic) with an exceptionally long period from May 2010 to March 2011 in which
252 relatively high *Dinophysis* abundances were recorded (around 2,000 cells·L⁻¹ of *D.*
253 *fortii*, Francé et al., 2018). These high abundances, never recorded again, were related
254 to long-lasting low salinity and extremely high temperatures in June – July surface
255 waters (<30 °C) causing a marked water column stratification (Francé et al., 2018).
256 High levels of okadaic acid (OA) and/or dinophysistoxin (DTX) in several instances
257 also led to halt shellfish harvesting along the French (Belin et al., 2020) and Spanish
258 coasts of the MS (García-Altarets et al., 2016; Fernández et al., 2019). Recurrent toxic
259 *Dinophysis* blooms have been recorded in the Thermaikos Gulf (Greece, North
260 Aegean Sea) since 2000, when they caused great economic losses (EU 5 million) to
261 aquaculture (Koukaras and Nikolaidis, 2004). More occasionally, high levels of DSP

262 toxins have been reported from the eastern Mediterranean (Orhanović et al., 1996;
263 Bazzoni et al., 2018) and Tunisian waters (Armi et al., 2012).
264 Nonetheless, there have been just a few cases of DSP diagnoses in humans, in the
265 Adriatic (Boni et al., 1992) and Tyrrhenian Seas (Lugliè et al., 2011), and two major
266 accidents. One occurred in 2000, when 200 people were hospitalized following the
267 above-mentioned *Dinophysis* bloom in the Thermaikos Gulf (Koukaras and
268 Nikolaidis, 2004). The other happened in 2010 in Piemonte (north-western Italy), with
269 more than 150 people harmed by the consumption of toxic mussels from the northern
270 Adriatic Sea (Pistocchi et al., 2012).
271 Other DSP producers widely distributed in the MS are two benthic species of the
272 genus *Prorocentrum* (Fig 3A), *P. lima* (Fig. 1F) and *P. rhathymum*, but no toxicity
273 events have been related with their presence.

274

275 2.2.2. Paralytic Shellfish Poisoning (PSP)

276 PSP events in the MS are related to toxins produced by species of the genus
277 *Alexandrium* and by *Gymnodinium catenatum*. Of the six *Alexandrium* species known
278 to produce PSP toxins found in the MS, *A. minutum*, the type species of the genus
279 (Fig. 1A), and *A. pacificum* (as *A. catenella* in records before 2014) are the most
280 commonly reported ones (Table 1, Fig. 3B). In some cases these species have reached
281 high densities (up to 10^7 cells·L⁻¹) causing seawater discolorations. *Alexandrium*
282 *taylorii* has also caused discolorations at several Spanish and Italian touristic places
283 (section 2.3.1, table S1) but no toxicity has ever been found in MS populations of this
284 species.

285 Reports of PSP events initially associated with *A. tamarense* (Boni et al., 1983;
286 Honsell et al., 1992; Abadie et al., 1999), a species that should not produce saxitoxins

287 (John et al., 2014), were later reinterpreted and attributed to *A. minutum* (Pistocchi et
288 al., 2012) or *A. pacificum* (Lilly et al., 2002). However, one strain of *A. tamarense*
289 from Sardinian coasts has recently been found to be toxic (Lugliè et al., 2017). Since
290 the first observations of massive natural fish mortalities in Egypt (Zaghloul and
291 Halim, 1992), *A. minutum* produced toxic blooms with consequent ban of both fishing
292 and shellfish harvesting in Morocco (Labib and Halim, 1995), Spain (Delgado et al.,
293 1990; Forteza et al., 1998), France (Belin et al., 2020) and Italy (Honsell et al., 1996).
294 After 2000, only a few cases of shellfish farm closures attributed to *A. minutum* have
295 been reported in northern Sardinia (Italy; Lugliè et al., 2011), Catalonia (Spain; Vila
296 et al., 2005; Bravo et al., 2008; Sampedro, 2018) and southern France coasts (Belin et
297 al., 2020). Because of a very similar non-toxic species discovered in the MS, *A.*
298 *tamutum*, the identification of *A. minutum* can be problematic and should be
299 confirmed by molecular or toxin analyses. *Alexandrium pacificum* was responsible for
300 toxic blooms along the Catalan coast (Bravo et al., 2008), in the Thau Lagoon
301 (Abadie et al., 1999), in Sardinia (Lugliè et al., 2011) and Sicily (Dell'Aversano et al.,
302 2019), at times causing shellfish harvesting closures (Vila et al., 2001; Bravo et al.,
303 2008).
304 *Alexandrium andersonii* and *A. ostenfeldii* are much less frequently recorded and
305 possibly overlooked or misidentified in plankton studies. At times their presence has
306 been traced as resting stages (e.g., Montresor et al., 1998; Bravo et al., 2006; Satta et
307 al., 2013). Two other *Alexandrium* species recorded in the MS, *A. balechii* and *A.*
308 *pseudogonyaulax*, do not produce PSP toxins but are considered potentially
309 ichthyotoxic.
310 *Gymnodinium catenatum* was first reported in southern Spain in 1987 (Bravo et al.,
311 1989). The worst, and apparently unique, fatal case of human intoxication in the

312 whole Mediterranean was due to a bloom of this species that caused 4 deaths and the
313 hospitalization of 23 people in Morocco in 1994 (Tagmouti-Talha et al., 1996).
314 Shellfish harvesting ban due to high concentrations of *G. catenatum* have however
315 been frequent in Andalusia (Spain) during the last 3 decades (HAEDAT). Records of
316 this species in the central and eastern MS should be considered with caution because
317 of possible misidentification of *G. impudicum* (Gómez, 2003).

318

319 2.2.3 Amnesic Shellfish Poisoning (ASP)

320 Sixteen of the 26 *Pseudo-nitzschia* species known to produce domoic acid (DA) have
321 been found so far in the MS. Species-level identification is problematic in light
322 microscopy and often requires the use of electron microscopy and/or molecular
323 markers. It follows that in most publications only the genus is reported, or taxa are
324 clustered into two ‘groups’, only distinguishing the thin (*P. delicatissima*-group) and
325 the thicker morphotypes (*P. seriata*-group). In the last decades, potentially toxic
326 *Pseudo-nitzschia* species have been identified properly from several locations of the
327 MS (Fig. 3C) where the presence of the cold-water species *Pseudo-nitzschia seriata*,
328 often reported in old studies, has never been confirmed.

329 Seasonal blooms of *Pseudo-nitzschia* spp., at times including toxic ones, occur all
330 along Mediterranean coasts (Fig. S1), with abundances up to several million cells·L⁻¹
331 (e.g., Caroppo et al., 2005; Cerino et al., 2005; Quiroga et al., 2006; Quijano-Scheggia
332 et al., 2008; Ljubešić et al., 2011; Marić et al., 2011; Cabrini et al., 2012; Ruggiero et
333 al., 2015; Taş and Lundholm, 2017; Totti et al., 2019a). Nevertheless, the detection of
334 DA has caused the closure of aquaculture plants only in a limited number of cases
335 (4% of toxicity events in HAEDAT) in southern Spain (HAEDAT) and France (Amzil
336 et al., 2001), whereas DA values below the regulatory limit have occasionally been

337 found in shellfish from the Adriatic Sea (Ciminiello et al., 2005; Ujević et al., 2010;
338 Arapov et al., 2016), Greece (Kaniou-Grigoriadou et al., 2005), and in 65% of 180
339 mussel samples from mid-Tyrrhenian waters (Rossi et al., 2016). In a few cases, the
340 presence of DA in bivalves was related to a specific taxon, i.e., *P. calliantha* along the
341 Croatian coast (Marić et al., 2011) and in the Gulf of Trieste (Honsell et al., 2008) and
342 *P. brasiliiana* in the Bizerte Lagoon in Tunisia (Sahraoui et al., 2011).
343 *Nitzschia bizertensis*, described from the Bizerte Lagoon (Tunisia), is one of the two
344 *Nitzschia* species known to produce domoic acid. At least in one case, the presence of
345 this species was related to the detection of domoic acid in mussels (Bouchouicha-
346 Smida et al., 2014). Less clear is the toxicity and the distribution of the other benthic
347 species *Halamphora coffeaeformis*.

348

349 2.2.4. *Ostreopsis* and species responsible of Ciguatera Fish Poisoning (CFP)

350 The benthic dinoflagellate *Ostreopsis* cf. *ovata* produces ovatoxins, which are
351 palytoxins-like molecules that can intoxicate humans by inhalation or ingestion of
352 contaminated seafood. The species was first detected in the MS in the plankton of
353 Villefranche-sur-Mer (France) after a strong *mistral* wind event in 1972 (Max Taylor,
354 pers. comm.), when it was identified with the name of the only species known at that
355 time, *O. siamensis*. The presence of the species was then documented from the coasts
356 of Lebanon in 1980 (Abboud-Abi Saab, 1989) and central Italy in 1986 (Zingone in
357 Tognetto et al., 1995). Around the 2000s, monitoring programs implemented
358 following a series of harmful events (see below) made it evident that *Ostreopsis*
359 species were growing all along the rocky shores of the northern MS (Fig. 4A) in
360 summer/autumn, thriving as epiphyte on macroalgae or epibiontic on a number of
361 benthic substrata, with concentrations up to 10^6 cells·g⁻¹ fresh weight of macroalgal

362 thalli (Mangialajo et al., 2011). At lower concentrations *Ostreopsis* spp. were also
363 found along the northern African coasts (Illoul et al., 2012; Ben Gharbia et al., 2019).
364 Of the three species so far identified in the MS, the most common and widespread is
365 *O. cf. ovata*, whereas *O. cf. siamensis* and *O. fattorussoi* have a much more restricted
366 distribution (Fig. 1E). An interesting aspect of the annual dynamics of *Ostreopsis*
367 species is the rather repetitive patterns of summer and/or autumn peaks, with timing
368 that vary from place to place and is scarcely related to temperature or to other obvious
369 environmental parameters (Zingone, 2010; Accoroni and Totti, 2016).

370 First problems caused by *Ostreopsis* in the MS were fish and invertebrate kills in
371 1998 along the coasts of Tuscany (northern Tyrrhenian Sea) (Sansoni et al., 2003;
372 Simoni et al., 2003). Some years later (2002) more than 200 people coming from the
373 beach of the city of Genoa (Ligurian Sea) were hospitalized with fever, red eyes and
374 wheeze (Ciminiello et al., 2006). The only known problems caused by benthic
375 microalgae at that time were those related to ciguatera fish poisoning (CFP) in
376 subtropical areas, whereas cases of toxic aerosol were only known for planktonic
377 *Karenia brevis* blooms in the Gulf of Mexico. In those years, similar human health
378 problems and dermatitis cases were reported from the Catalonia and Balearic Islands
379 (Vila et al., 2008), French (Cohu et al., 2013) and Algerian coasts (Illoul et al., 2012),
380 and are still reported nowadays at several MS places (e.g., Croatian coast, Ninčević
381 Gladan et al., 2019). Both the presence of toxins in the aerosol (Ciminiello et al.,
382 2014) and toxicological data on the effects of inhalation exposure in mice (Poli et al.,
383 2018) support a link between *Ostreopsis* toxins and the respiratory symptoms reported
384 during blooms. However, those health problems do not occur during all phases of a
385 bloom (Vila et al., 2016) and are quite sporadic compared to the widespread and often
386 massive presence of the suspected causative species.

387 The presence of *Ostreopsis* toxins in marine animals used as food and their impacts
388 on the animal health are relevant for their sanitary implications, which are still
389 controversial (Tubaro et al., 2011). Apparently healthy organisms (e.g., mussels and
390 sea urchins) during *Ostreopsis* blooms can accumulate fairly large amount of toxins
391 (Aligizaki et al., 2008; E. Fattorusso & V. Soprano, pers. comm.), but macroscopic
392 damages have been reported for various benthic organisms in the MS (Sansoni et al.,
393 2003, Simoni et al., 2003; Accoroni and Totti, 2016) and elsewhere (Shears and Ross,
394 2009). In mussels, *Ostreopsis* can induce important and not completely reversible
395 ultrastructural damages (Carella et al., 2015) and immunological, histological and
396 oxidative responses (Gorbi et al., 2013) while in sea urchins *Ostreopsis* blooms affect
397 reproduction and offspring health (Migliaccio et al., 2016).

398 Four species of the dinoflagellate genus *Gambierdiscus*, which can produce CFP
399 toxins, have recently been found in the MS. *Gambierdiscus australes*, *G. cf.*
400 *belizeanus*, *G. carolinianus*, *G. silvae* and some unidentified *Gambierdiscus* spp.,
401 have been reported from the Balearic Islands (Tudó et al., 2018), Greece and Cyprus
402 (Aligizaki and Nikolaidis, 2008; Holland et al., 2013; Aligizaki et al., 2018; Tudó et
403 al., 2018), with the highest diversity in Crete. *Fukuyoa paulensis* also has been found
404 in the Balearic Islands (Laza-Martínez et al., 2016) and Cyprus (Tudó et al., 2018).
405 Yet CFP cases are not known in the MS countries with the exception of a suspected
406 case of ciguatoxins in rabbitfish (*Siganus rivolutus*) reported from Israeli coasts
407 (Bentur and Spanier, 2007).

408

409 2.2.5. Azaspiracid Shellfish Poisoning (AZP)

410 The toxins azaspiracids (AZAs), produced by a number of dinoflagellate species of
411 the genera *Azadinium* and *Amphidoma*, and the human syndrome they can cause,

412 AZP, have been discovered at the beginning of this century (James et al., 2002).
413 Subsequently AZAs have been reported in shellfish from numerous sites, including
414 the MS (Bacchiocchi et al., 2015). A new species described from the MS, *A.*
415 *dexteroporum* (Percopo et al., 2013, Fig. 1 B), produces a whole suite of AZAs that
416 can cause direct harm to molluscs (Rossi et al., 2017; Giuliani et al., 2019). Another
417 toxic *Azadinium*, *A. poporum*, has been found in Greek waters (Luo et al., 2018) but
418 no impacts related to AZAs have been reported so far.

419

420 2.2.6 Ichthyotoxicity

421 About half of the potentially toxic MS species produce a variety of toxins that differ
422 from those related to the syndromes mentioned in the previous sections. Of these, the
423 majority (28 species, Table 1) produce substances that have been associated with fish
424 and/or shellfish kills. With a few exceptions, species in this list are unarmoured
425 dinoflagellates, e.g., *Karenia* and *Karlodinium*, and other flagellates belonging to the
426 prymnesiophytes, raphidophytes and dictyochophytes, which are all hardly
427 identifiable in fixed material under the light microscope, and hence are overlooked in
428 most monitoring and ecological investigations. The large majority of the information
429 on the presence of these ichthyotoxic species (Fig. 4B) comes from fish mortality
430 events, mainly located near fish-farming plants, in which the identification of the
431 culprit became necessary.

432 The few fish mortality events in the MS known before 1975 were related to HB-
433 HABs of non-ichthyotoxic species causing anoxia in bottom waters (see section 2.3.1)
434 rather than to ichthyotoxic species (Jacques and Sournia, 1978-79). In the subsequent
435 years, fish kills by ichthyotoxic species were reported sporadically from Catalan
436 coasts, Spain (Garcés et al., 1999), caused by *Karlodinium* spp., and Sardinia (Italy),

437 caused by *Chattonella subsalsa* (Stacca et al., 2016). Occasional fish mortality events
438 were related to *Prymnesium* spp., in the Ebro Delta (Spain, Comín and Ferrer, 1978)
439 and in a Tuscany lagoon (Italy, Mattioli and Simoni, 1999), *Karenia selliformis* in the
440 Gulf of Gabes (Tunisia, Romdhane et al., 1998; Feki et al., 2013) and *Karenia brevis*
441 and *Pseudochattonella* cf. *verruculosa* in Greece (Ignatiades and Gotsis-Skretas,
442 2010). In other cases, fish kills occurred during blooms of species toxic to humans,
443 like in Egypt in 1987 (Zaghloul and Halim, 1992; Labid and Halim, 1995) where
444 *Alexandrium minutum* was the culprit. No fish or shellfish kill accidents in the MS
445 have ever been associated with blooms of two potentially ichthyotoxic *Alexandrium*
446 species, *A. balechii* and *A. pseudogonyaulax*.

447 Benthic cyanobacteria are poorly investigated in Mediterranean waters, but blooms of
448 filamentous cyanobacteria have been the cause of massive fish mortalities in
449 Alexandria waters (Egypt) during spring 2005 (Ismael, 2012).

450

451 2.2.7 Other toxins

452 The dinoflagellates *Gonyaulax spinifera*, *Lingulodinium polyedra* and *Protoceratium*
453 *reticulatum*, which are quite widespread in the MS (Fig. 4 B), produce yessotoxins
454 (YTX). These substances were initially associated to DSP because their presence
455 gives similar positive results in mouse bioassay, but they are not considered toxic to
456 humans (Tubaro et al., 2010). However, YTXs caused economic impacts in 2002,
457 2004 and 2007, when mussel harvesting was halted for a long time (average closure
458 153 days) in the north-western Adriatic Sea (Poletti et al., 2008).

459 *Vulcanodinium rugosum* produces pinnatoxins (Rhodes et al., 2010; Nézan and
460 Chomérat, 2011) a neurotoxin that has lethal effects on sea urchin larvae, oysters and

461 *Artemia*. Currently there are no problems related to this species, while toxic effects on
462 humans are not known.

463

464 **2.3. Non-toxic events**

465 Independent from toxin production, all microalgae may exert a negative impact when
466 they reach a high biomass producing seawater discolouration, mucilages or anoxia in
467 bottom waters (Zingone and Enevoldsen, 2000). Although several microalgal species
468 are frequently associated with these HB-HABs, as detailed in the next sections, the
469 number of species that may cause harm with no specific toxin production is in theory
470 unlimited, and can vary from place to place. For this reason it is not possible to define
471 a global or regional list of non-toxic harmful microalgae. In addition to HB-HAB-
472 formers, some non-toxic species, mainly diatoms, may cause mechanical harm to
473 invertebrates' gills (Bell, 1961), but no information on such events is available for the
474 MS. In case of fish or invertebrate kills, at time it is hard to discern whether the cause
475 has been anoxia, toxic substances or mechanical damages. In many cases, species
476 known to produce toxins may produce non-toxic HB-HABs, which have no impact on
477 human or marine fauna health but important consequences for tourism. For all these
478 reasons, the boundaries between events described in the previous and next sections
479 cannot always be well defined.

480

481 *2.3.1 Discolorations*

482 In the MS, discolouration or anoxia have frequently been caused by unarmoured
483 dinoflagellates either toxic (e.g., *Margalefidinium polykrikoides*) or non-toxic (e.g.,
484 *Noctiluca scintillans*), but also by numerous armored dinoflagellates, diatoms,
485 prasinophytes, prymnesiophytes and raphidophytes (Table S1). Change of seawater

486 color caused by HB-HABs (Fig. 5) have been noticed since the first half of the XX
487 century in both lagoons and coastal sites, where they were given several names (*purga*
488 *de mar*, *punti verdi*) before the one of *red tides* gained popularity. The oldest records
489 include discolorations caused by *Chattonella subsalsa* in 1956 in the Algiers harbor
490 (Hollande and Enjumet, 1957), *Alexandrium minutum* in 1957 in the Alexandria
491 harbor (Halim, 1960) and *Prorocentrum cordatum* in the Gulf of Naples in September
492 1962 (Yamazi, 1964).

493 Different dinoflagellates (e.g., *Alexandrium* spp., *Noctiluca scintillans*, *Karlodinium*
494 spp.), raphidophytes (*Chattonella subsalsa* and *Fibrocapsa japonica*, Fig. 1D) and
495 chlorophytes (*Tetraselmis wettsteinii* and *Pyramimonas* spp.) occasionally produced
496 discoloration (Table S1, Fig. 5), which in some cases were also associated with fish
497 kills and/or massive death of marine invertebrates caused by anoxic conditions (e.g.,
498 Arzul, 1994; Halim and Labib, 1996; Garcés et al., 1999). A couple of such cases of
499 fish mortality events attributed to anoxia were already reported in the review by
500 Jacques and Sournia (1978-79): in Ismir Bay (Nümann, 1955, in Jacques and Sournia,
501 1978-79) and in the Adriatic Sea (Piccinetti and Manfrin, 1969; Frogliani, 1970), during
502 blooms of *Gymnodinium* sp. and *Protoperdinium depressum*, respectively.

503 Discolorations were particularly frequent in the northern Adriatic Sea in summer in
504 the 1970–'80s, when dinoflagellate blooms (e.g., *Lingulodinium polyedra*,
505 *Alexandrium mediterraneum* and *Lepidodinium chlorophorum*) turned the sea into
506 various colours (Boni, 1983, Table S1), at times extending offshore as in the case of
507 *N. scintillans* in 1980 (Fonda Umani et al., 2004) and *L. chlorophorum* in 1984
508 (Artegiani et al., 1985). Some summer blooms were caused by diatoms (e.g.,
509 *Skeletonema marinoi* and *Chaetoceros* spp.), particularly after intense freshwater
510 inputs (Boni, 1983; Regione Emilia Romagna, 1982-2018). Over the last decades

511 blooms of *F. japonica* (Fig 1D) became common in late summer (Cucchiari et al.,
512 2008) in shallow coastal waters where they lasted up to 20–40 days. Along the eastern
513 Adriatic coast, ‘red tides’ were limited to eutrophicated semi-enclosed bays
514 (Marasović et al., 1991) or to unusual phenomena such as bloom of the silicoflagellate
515 *Octactis* (formerly *Distephanus*) *speculum* in summer 1983 in bottom waters in the
516 Gulf of Trieste, causing anoxia (Fanuko, 1989). An increasing number of
517 discolorations have been observed over two decades in the Golden Horn Estuary of
518 the Sea of Marmara (Taş et al., 2016). An unusual bloom of the coccolithophore
519 *Holococcolithophora sphaeroidea* (as *Calyptrosphaera sphaeroidea*) caused a white-
520 green-turquoise discoloration in a vast area off the Tarragona harbor (Spain, Cros et
521 al., 2002). The most recent event has been a long-lasting bloom of *Margalefidinium*
522 cf. *polykrikoides* that produced a yellow brownish discoloration in a touristic area of
523 the Ionian Sea (Italy) in July-August 2018, recurring in the same place in summer
524 2019 (Roselli et al., 2020).

525 In summer, discolorations can be a serious problem along Mediterranean beaches
526 where they have an impact on tourism and recreational use of the sea. This is the case
527 of the recurrent *Alexandrium taylorii* blooms along the Sicilian and Sardinian coasts
528 (Italy) and in the Balearic Islands (Spain) (e.g., Basterretxea et al., 2005; Giacobbe et
529 al., 2007; Satta et al., 2010; Sampedro, 2018).

530

531 2.3.2 Mucilages

532 In the MS, a number of cases of mucilaginous aggregate formation related to
533 microalgal growth have been described, the most conspicuous of which occurred in
534 the northern Adriatic Sea in the 1990s. Mucilaginous macroaggregates represent the
535 last stage of aggregation of organic matter, mainly refractory polysaccharides derived

536 from phytoplankton exudates (Myklestad, 1995) and/or from bacterial capsular
537 material (Stoderegger and Herndl, 1998) whose hydrolysis cannot be sustained by
538 phosphorous-limited bacteria (Danovaro et al., 2005). Whereas marine snow
539 (aggregates of 0.5-1 cm diameter) is common in all the oceans (Simon et al., 2002),
540 the mucilage event in the northern Adriatic Sea was unique in that those aggregates
541 covered hundred square kilometres of both coastal and offshore areas. The formation
542 of larger aggregates was favored by the strong stratification of the water column and
543 reduced circulation that retained freshwater in the northern Adriatic basin (Russo et
544 al., 2005). The direct responsible of the phenomenon were often thought to be the
545 most abundant phytoplankton species in the aggregates, such as *Cylindrotheca*
546 *closterium* (Revelante and Gilmartin, 1991) and *Gonyaulax fragilis* (Pompei et al.,
547 2003), both capable to produce large amounts of refractory polysaccharides (Pistocchi
548 et al., 2005; Urbani et al., 2005). In fact, phytoplankton communities associated with
549 mucilage aggregates largely vary, depending on sampling area and period (Totti et al.,
550 2005, and references therein), while the aggregates represent a self-sustained
551 microcosm hosting a rich microorganism community (Simon et al., 2002).
552 Pelagic mucilages have been reported at several other Mediterranean sites, such as the
553 Greek (Gotsis-Skretas, 1995; Nikolaidis et al., 2008) and Catalan coasts (Sampedro et
554 al., 2007) where *Gonyaulax fragilis* was thought to be involved in their production,
555 and the Sea of Marmara (Turkey) where *Cylindrotheca closterium*, *Skeletonema*
556 *costatum* and *Gonyaulax fragilis* were indicated as the most abundant species
557 (Tüfekçi et al., 2010). In the Tyrrhenian Sea, extensive pelagic aggregates were
558 observed in 1991, 2000 and 2012 (Fig. 5 A, Calvo et al., 1991; Innamorati et al.,
559 1993; Escalera et al., 2018). Foam accumulated massively along the Catalan coast in

560 March 2006 during a *Phaeocystis* sp. bloom, an event that was related to anomalous
561 hydrographic winter conditions (Arin et al., 2014).

562 Massive mucilage events have also concerned the benthic environment. *Ostreopsis* cf.
563 *ovata* during intense blooms forms a network-shaped mucilaginous biofilm that can
564 harm benthic invertebrates (Schiaparelli et al., 2007). In the Tyrrhenian and Ligurian
565 Seas (western MS), benthic mucilages have occurred since 1991 (Sartoni and Sonni,
566 1991), and have been attributed to the massive growth of several macro- and
567 microalgae such as the filamentous brown alga *Acinetospora crinita* and the colonial
568 pelagophytes *Nematochryopsis marina* and *Chrysonephos lewisii* (Giuliani et al.,
569 2005; Schiaparelli et al., 2007). The allochthonous pelagophyte *Chrysophaeum*
570 *taylorii*, recorded in the western MS since 2005, in recent years was involved in the
571 formation of dense layers of mucous covering macroalgae, gorgonians and the
572 surrounding rocks (Lugliè et al., 2008; Caronni et al., 2015).

573

574 **3. Trends in the Mediterranean HABs**

575 *3.1 General trends*

576 The MS has undergone profound changes over the last centuries. Human action has
577 mainly been visible along the coasts of the basin, which have become increasingly
578 populated and deeply modified by coastal and riverine engineering and deforestation
579 which, along with cultural eutrophication, are all potential drivers of deep changes in
580 phytoplankton communities (Garcés and Camp, 2012). Natural and/or man-induced
581 meteorological and climatic variations superimpose to these changes often with an
582 amplifying effect. The most striking characteristic of the MS HABs over the last 50
583 yrs, which approximately correspond to the time since when they have been studied
584 more intensively, is the remarkable increase of the toxic species list, from a few taxa

585 to the more than 80 of the present review (Fig. 2). Over the same period, the records
586 of these species across the MS have also remarkably increased (Fig. 6). This trend is
587 parallel to that of the increased list of toxic species and of their records worldwide,
588 which is an obvious result of the intensification of the taxonomic and toxin studies on
589 marine microalgae (Zingone et al., 2017). The increase of the records of actual HAB
590 events from the less than 30 cases listed by Jacques and Sournia (1978-1979) and
591 Honsell et al. (1995) to the several hundred cases of halted aquaculture operations,
592 seawater discoloration and minor human health accidents presently recorded in
593 HAEDAT is also impressive (Fig. 7). Damages to aquaculture caused by ASP and
594 PSP toxins in mussels have been limited over the last 30 years while DSP cases have
595 represented about 75% of the harmful events, with an increase between the decade
596 1987-1997 and the two following ones (Fig. 7). This trend should however be
597 interpreted with caution because it has been paralleled by a remarkable growth of the
598 coastal MS population (section 1), much more intensive use of marine resources, and
599 consequent raise of the level of attention to the integrity and safety of marine
600 resources.

601 In fact, toxic blooms as well as mucilage events and discolorations in the MS have
602 generally shown an unpredictable interannual periodicity, like in the case of the
603 conspicuous blooms of *Noctiluca scintillans* in the Adriatic Sea (Fonda Umani et al.,
604 2004), Moroccan (Tahri Joutei et al., 2003), Catalan (Lopez and Arte, 1971) and
605 French coasts (M.-O. Soyer in Jacques and Sournia, 1978-1979). There are cases of
606 decreases, e.g., the blooms of *Alexandrium pacificum* occurring on the Catalan coast
607 from 1996 to 1998 (Vila et al., 2001) but rarely recorded afterwards (Sampedro,
608 2018). Blooms of *A. minutum* were recurrent in Egyptian waters but not recorded any
609 longer after 1994 (Ismael and Halim, 2001), while their frequency doubled from 2000

610 to 2012 along the Catalan coast (Sampedro, 2018). Blooms of the ciliate *Mesodinium*
611 *rubrum* hosting cryptophyte chloroplasts were not recorded in the MS (Jacques and
612 Sournia, 1978-1979) until their occurrence in both the Adriatic (Sorokin and
613 Ravagnan, 1999) and Tyrrhenian Seas (Siano et al., 2006), and afterwards have only
614 been observed in 2017 in the North Aegean Sea (Genitsaris et al., 2019).

615 In the case of *Ostreopsis* cf. *ovata*, rather than an increase the phenomenon in the MS
616 has shown a sudden upsurge around the 2000, followed by an expansion of the known
617 range for the species in the next years and a relative stability in the following decade.
618 Indeed *Ostreopsis* cf. *ovata* provides the most evident case of range expansion and
619 increased impact over time in the MS. Although benthic microalgae have received
620 scarce attention until the late 20th century, it is unlikely that the species might have
621 been abundant but undetected before. The apparent sudden range expansion and
622 impact of *Ostreopsis* cf. *ovata* is in line with an increasing trend of species of the
623 same genus in New Zealand and some other temperate areas around the world
624 (Parsons et al., 2012). On the other hand, no clear increase of the impact or of species
625 abundance has been reported since the 2000 outburst, while the above-mentioned
626 range expansion has coincided with a dramatic increase in monitoring programs and
627 research projects focused on benthic microalgae. Initially, the sudden relevance of the
628 phenomenon was associated with an increase of temperature in the MS, based on the
629 belief that all *Ostreopsis* species were of tropical origin. In fact, *Ostreopsis* cf. *ovata*
630 and its close relatives are widely distributed in temperate areas, also matching the
631 apparent preference of the species for moderately high rather than very high
632 temperature (Mangialajo et al., 2011; Scalco et al., 2012). Overall, the trend observed
633 for this species in the MS, with an outburst followed by a stabilizing trend, recalls that
634 of an invasive species rather than that of a species favored by a temperature increase.

635

636 *3.2 HAB trends in the Adriatic Sea, a case study*

637 The Adriatic Sea (AS) represents a unique system for its semi-enclosed morphology,
638 shallow depth and oligotrophic nature in most parts but with eutrophic characteristics
639 along the north-western coasts driven by inputs from the Po River and other rivers
640 (Mozetič et al., 2010; Cozzi and Giani, 2011). The AS is considered one of the
641 hotspots of MS HABs (Garcés and Camp, 2012), in terms of both occurrence and
642 impacts. However, compared to the great variety of potentially toxic species (Mozetič
643 et al., 2019), toxicity cases are limited, and the most common toxins found above the
644 regulatory limits in the Adriatic shellfish to date are DSP toxins (okadaic acid group)
645 and other lipophilic toxins (yessotoxins and pectenotoxins).

646 Because of the early development of marine-related activities, there is a wealth of
647 information from the area dating back to the last century, which allows some insights
648 on possible HAB trends. Phytoplankton in certain areas of the AS (e.g., Gulf of
649 Trieste, Gulf of Venice, Senigallia-Susak transect, Kaštela Bay) have been
650 extensively studied for decades (Ninčević-Gladan et al., 2010; Bernardi Aubry et al.,
651 2012; Marić et al., 2012; Mozetič et al., 2012; Cerino et al., 2019; Totti et al., 2019a),
652 highlighting a number of changes, such as trends or regime shifts in main
653 phytoplankton groups (Mozetič et al., 2010; Totti et al., 2019a) and in bloom forming
654 species (Cabrini et al., 2012). However, no trends specifically related to toxic species
655 is evident from these long-term studies, neither in terms of increased frequency nor of
656 abundance. In fact, most studies on HAB species are snapshots of isolated toxic
657 episodes (Pistocchi et al., 2012, and references therein). Similar conclusions can be
658 drawn also from toxicity events: aquaculture operations have been halted frequently
659 over the last 20 years (section 2.2.1), but without any significant trend for DSP events.

660 Nevertheless, some changes in phytoplankton community structure of the AS have
661 involved a number of HAB species, such as *Pseudo-nitzschia multistriata*, an
662 allochthonous species (section 2.1) that became a regular component of the autumn
663 phytoplankton communities of the NW AS (Totti et al., 2019a). In the Gulf of Trieste,
664 previously rare *Dinophysis tripos* have become a regular member of the autumn
665 phytoplankton assemblages since 2010, along with higher temperatures recorded in
666 this decade (Francé et al., 2018), whereas further south *D. sacculus* has replaced *D.*
667 *caudata* as one of the indicator species of spring phytoplankton communities (Totti et
668 al., 2019a).

669 HB-HABs caused by dinoflagellates, occurring in summer and often associated with
670 water discoloration and bottom anoxia, were a major problem in the AS until the end
671 of the 1980s (see section 2.3.1). At the time, because of the heavy impact on the local
672 economy, the Italian government adopted countermeasures to reduce P content in
673 detergents and improve the urban wastewater treatment plants, leading to a strong
674 reduction of P load in coastal waters. Since the end of the 1980s, summer
675 dinoflagellate blooms became a rarer phenomenon, their decline coinciding with the
676 years of large mucilaginous macroaggregate appearance.

677 Mucilages in the AS (see section 2.3.2) were known since the beginning of 1700,
678 when they were named ‘mare sporco’. In more recent years, massive episodes have
679 occurred in the years 1988 to 1991 and 1997 to 2004, typically in summer (Giani et
680 al., 2005), while a spatial and temporal reduction occurred in subsequent years. An
681 anomalous occurrence in autumn-winter was reported in 2006-2007, probably in
682 relation to a water temperature increase (Danovaro et al., 2009). The mucilage
683 appearance, and the concurrent disappearance of summer water discolorations have
684 both been associated with the decrease of inorganic and organic P (Degobbis et al.,

685 2005), but also to hydrographic changes related to large-scale climatic changes around
686 the end of the '80s, which could have driven a regime-shift affecting not only the AS
687 but also other European Seas (Conversi et al., 2010).

688 In the last decade (2008-2018), HB-HABs of both diatoms and dinoflagellates
689 occurred without a regular temporal pattern, reflecting the meteorological events that
690 nowadays tend to be more intense and unhampered by a regular seasonal rhythm
691 (Totti et al., 2019a, b). Blooms of *Fibrocapsa japonica* that were common at the end
692 of the 1990s seem to be rarer since 2012 (Regione Emilia-Romagna, 1982-2018), and
693 mucilage events occurred shortly in 2014 and in 2018 (Regione Emilia-Romagna,
694 1982-2018).

695 As a whole, HABs in the AS show unpredictable time variability that is partly related
696 to the irregularity and intensity of meteorological events in the last decades.

697 Prolonged periods of drought (Cozzi et al., 2019) with oligotrophic conditions
698 (Mozetič et al., 2010) alternate with nutrient pulses from continental water runoff that
699 can drive the occurrence of anomalous intense blooms at any time of the year (Totti et
700 al., 2019a).

701

702 **4. Conclusions**

703 A deep knowledge on the spatial and temporal distribution of harmful species and the
704 blooms that they produce is an indispensable goal towards a safe use of marine
705 resources and an informed management and planning of the coastal zone. In the MS
706 this goal is even more crucial considering the importance of the economy deriving
707 from the use of the sea for tourism and recreational use, fishery and aquaculture. The
708 information about HABs has grown remarkably over the last 50 years since the first
709 review (Jacques and Sournia, 1978-1979) all over the MS areas. However, the marked

710 west-east and north-south gradients in the knowledge of HABs and HAB species
711 distribution persist, with long traits of coast with scarce or no information available.
712 Overall, the MS hosts a high number of potentially toxic species, many of which have
713 a wide distribution across its coastal waters. Yet the cases of intoxication are
714 extremely rare, while the impact on aquaculture appears to be limited to a few hot
715 spots in the northern Adriatic, Spain and France coasts. A variety of toxins have
716 actually been detected in several instances in microalgae strains from the MS, while
717 seafood toxicity, when detected, has commonly remained below the safety limits. The
718 typical oligotrophic offshore Mediterranean waters that influence most coastal areas
719 and the enhanced alongshore circulation in many places may play a role in keeping
720 toxic algae at levels rarely exceeding critical density thresholds, thus preventing their
721 excessive accumulation in seafood. On the other hand, quite effective monitoring
722 operations have accompanied the development of aquaculture over the last decades,
723 thus reducing the possibility of accidents to a minimum level.

724 In terms of microalgal toxins, the only major concern seems to reside in the large
725 amount of palytoxin-like substances that every summer accumulate along the rocky
726 Mediterranean shores because of *Ostreopsis* blooms. Although sea urchins and wild
727 mussels inhabiting those environments at time accumulate those toxins to
728 considerable levels, no cases of seafood intoxication have occurred so far.

729 Contaminated herbivorous fishes represent a problem in areas where they ingest
730 macroalgal substrates colonized by toxic microalgae, i.e. in the ciguatera areas, but
731 species capable of this transfer link may be missing in the MS trophic webs, or toxins
732 are neutralized in the transfer. Nonetheless, the guard level must be kept high because
733 sudden changes might occur, e.g., due to penetration of benthic herbivorous fish in the
734 MS and consequent novelties in the local food webs.

735 Overall, the present overview demonstrates a relatively low risk deriving from toxic
736 blooms and a higher risk from high biomass blooms affecting the aesthetic qualities of
737 coastal areas devoted to tourism in the MS. No clear trends in occurrence nor
738 expansions emerge for either toxic or HB-HABs. While EU regulation and national
739 initiatives have promoted actions addressing seawater quality and aiming at a good
740 environmental status (GES), human densities along the coasts is predicted to keep on
741 increasing in the next decades. Therefore, a larger use of marine resources in the
742 future, in the MS like in other coastal areas of the world, will probably lead to an
743 increased impact of the risks posed by HABs even in absence of any trends in their
744 abundance and frequency (Zingone and Wyatt, 2005). In addition, predicted changes
745 in climate and consequent modifications in hydrographical features may drive local
746 variations in microbial populations both in the plankton and in the benthos. Continued
747 monitoring and further studies on HAB patterns and trends are therefore mandatory
748 goals to be able to predict their evolution and protect human health and wellbeing in
749 the MS.

750

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1487 **Figure captions**

1488 **Figure 1:** Examples of toxic species from the Mediterranean Sea. A) *Alexandrium*
1489 *minutum* stained with calcofluor. B) *Azadinium dexteroporum*. C) *Dinophysis*
1490 *sacculus*. D) *Fibrocapsa japonica*. E) *Ostreopsis fattorussoi* stained with Calcofluor
1491 (courtesy of S. Accoroni). F) *Prorocentrum lima*. G) *Pseudo-nitzschia multistriata*.
1492 Scale bars in A and B: 5 μm ; in C, D, E, F and G: 20 μm .

1493 **Figure 2:** Cumulative numbers of known toxic species in the Mediterranean Sea in
1494 different years.

1495 **Figure 3:** Geographic range of potentially toxic species in the Mediterranean Sea.
1496 Distribution of species known to produce toxins related to: A) Diarrhetic Shellfish
1497 Poisoning (DSP), *Dinophysis* spp. and the benthic species *Prorocentrum lima* and P.

1498 *rhathymum*. B) Paralytic Shellfish Poisoning (PSP), *Alexandrium* spp. and
1499 *Gymnodinium catenatum*. C) Amnesic Shellfish Poisoning (ASP), *Pseudo-nitzschia*
1500 spp. and *Nitzschia bizertensis*. For the genera *Dinophysis*, *Pseudo-nitzschia* and
1501 *Alexandrium*, which include both toxic and non-toxic species, the maps represent only
1502 toxic species and, in case of cryptic or problematic species, only the records validated
1503 by electron microscopy, molecular methods and/or toxin production.

1504 **Figure 4:** Geographic range of potentially toxic species in the Mediterranean Sea. A)
1505 *Ostreopsis* spp. (mostly *O. cf. ovata*) and species related to the Ciguatera Fish
1506 Poisoning (CFP). B) Species producing ichthyotoxins (*Alexandrium*
1507 *pseudogonyaulax*, *Karenia* spp., *Karlodinium* spp., *Chattonella* spp., *Vicicitus*
1508 *globosus*, *Prymnesium* spp., etc.) and other toxins. The latter include mainly a few
1509 widespread dinoflagellate species that produce yessotoxins (*Lingulodinium polyedra*,
1510 *Gonyaulax spinifera* and *Protoceratium reticulatum*), but also other dinoflagellates
1511 producing azaspiracids (*Azadinum* spp.), pinnatoxins (*Vulcanodinium rugosum*) and
1512 other toxins with poorly known effects (e.g., *Prorocentrum* spp., *Margalefidinium*
1513 *polykrikoides*). See Table 1 for a complete list.

1514 **Figure 5:** A) Mat of *Oscillatoria acutissima* in the Eastern Harbour of Alexandria
1515 (Egypt). B) Bloom of *Noctiluca scintillans* in Thermaikos Gulf (Thessaloniki,
1516 Greece). C) Discoloration caused by *Euglena viridis* in the Golden Horn Estuary (Sea
1517 of Marmara, Turkey). D) Shellfish mortality in Ras El-Bar (Egypt) in 2011 due to the
1518 proliferation of *N. scintillans* and consequent oxygen depletion. E) Pelagic mucilages
1519 in the Gulf of Naples (Italy).

1520 **Figure 6:** Distribution of potentially toxic species, mucilages and discolorations in the
1521 Mediterranean Sea. A) Distribution of species known to be toxic and harmful events

1522 until 1995 as reported in Jacques and Sournia (1978-1979) and Honsell et al. (1995).

1523 B) Distribution of potentially toxic species (excluding *Ostreopsis* and CFP species)

1524 and harmful events updated to the present status of knowledge. The position of the

1525 circles in several cases has been slightly modified to reduce overlapping.

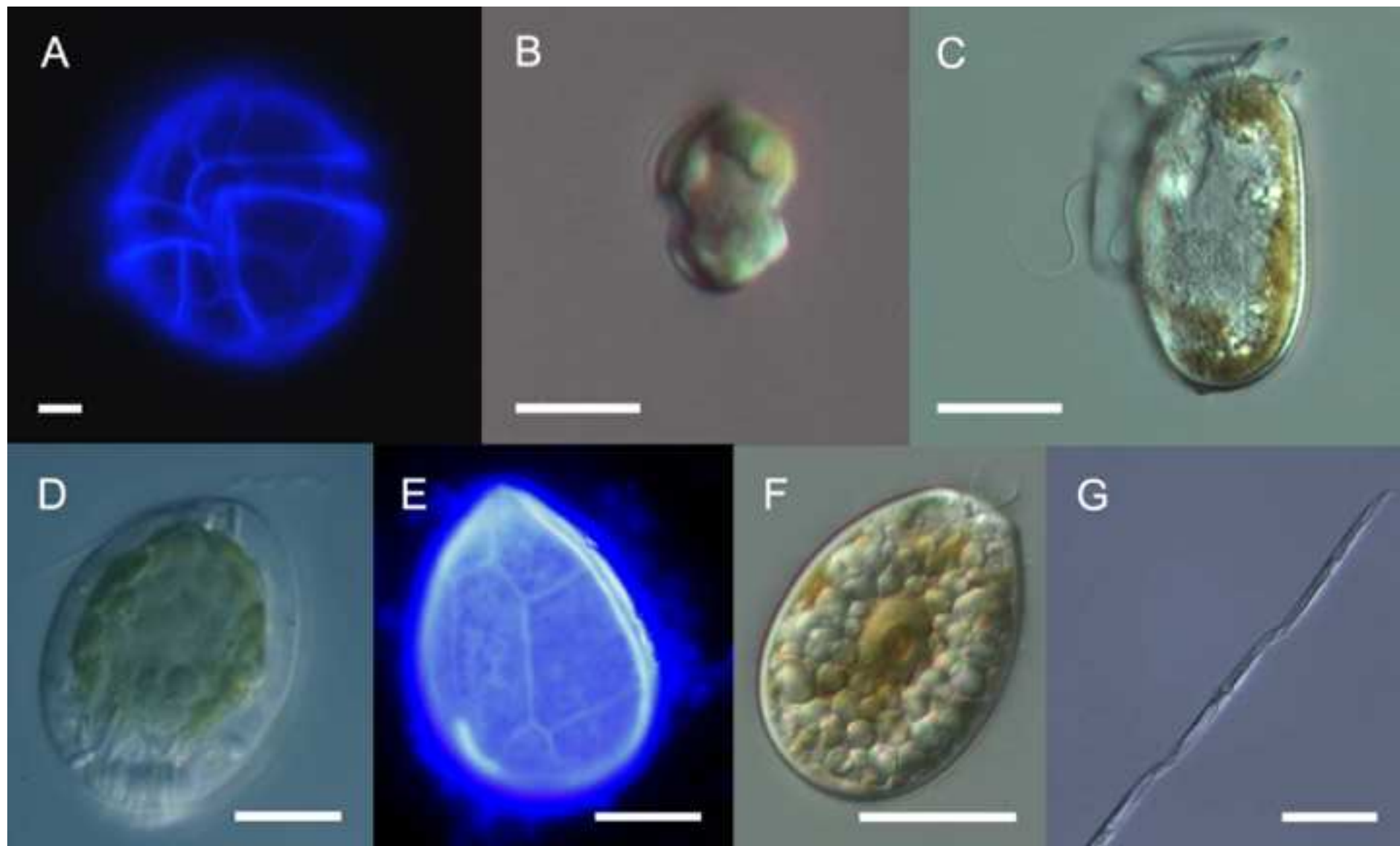
1526 **Figure 7.** Harmful events related to microalgae in the Mediterranean Sea (n=501)

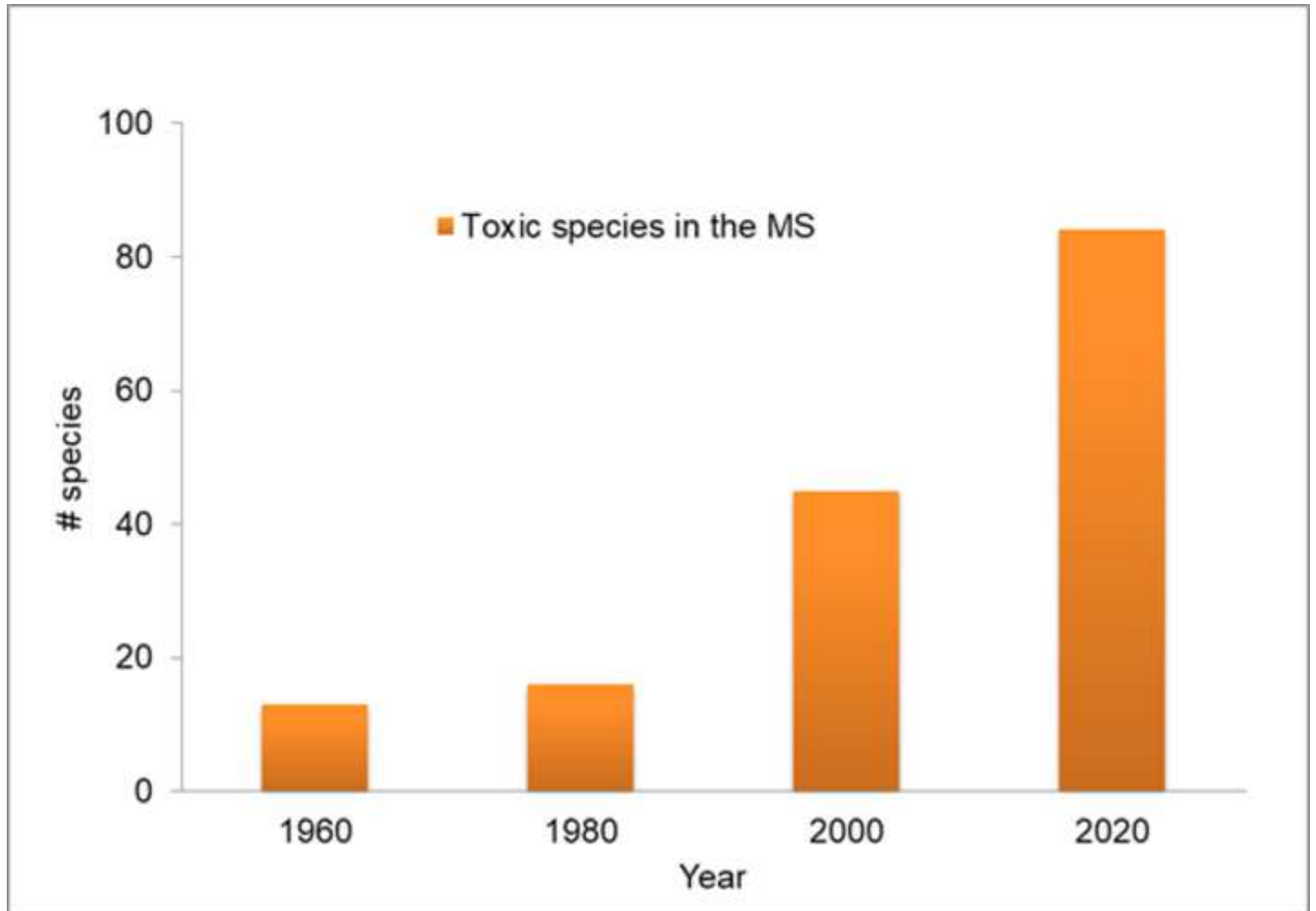
1527 based on records in the Harmful Algae Event Database HAEDAT

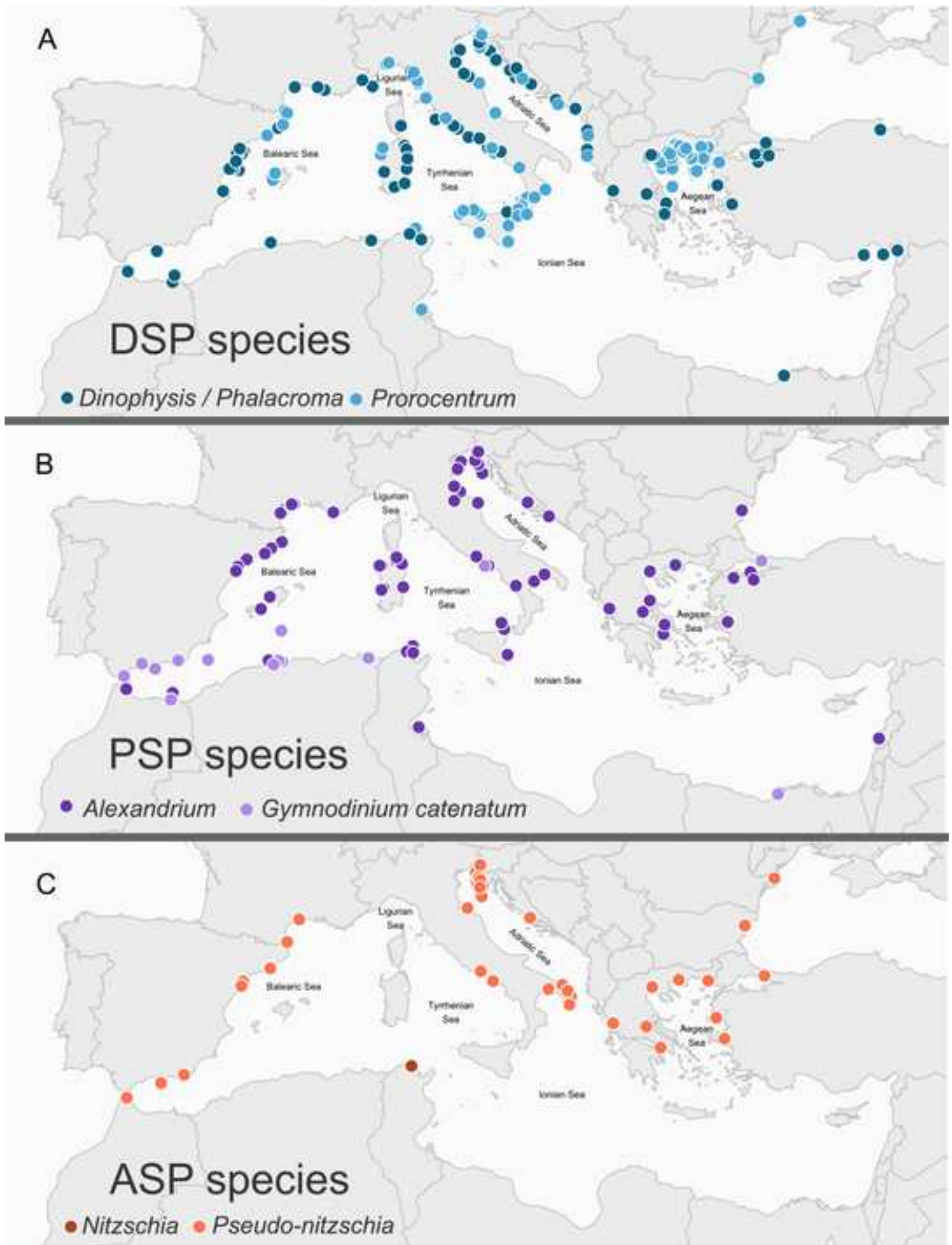
1528 (<http://haedat.iode.org/>). High density phytoplankton blooms with no impacts were

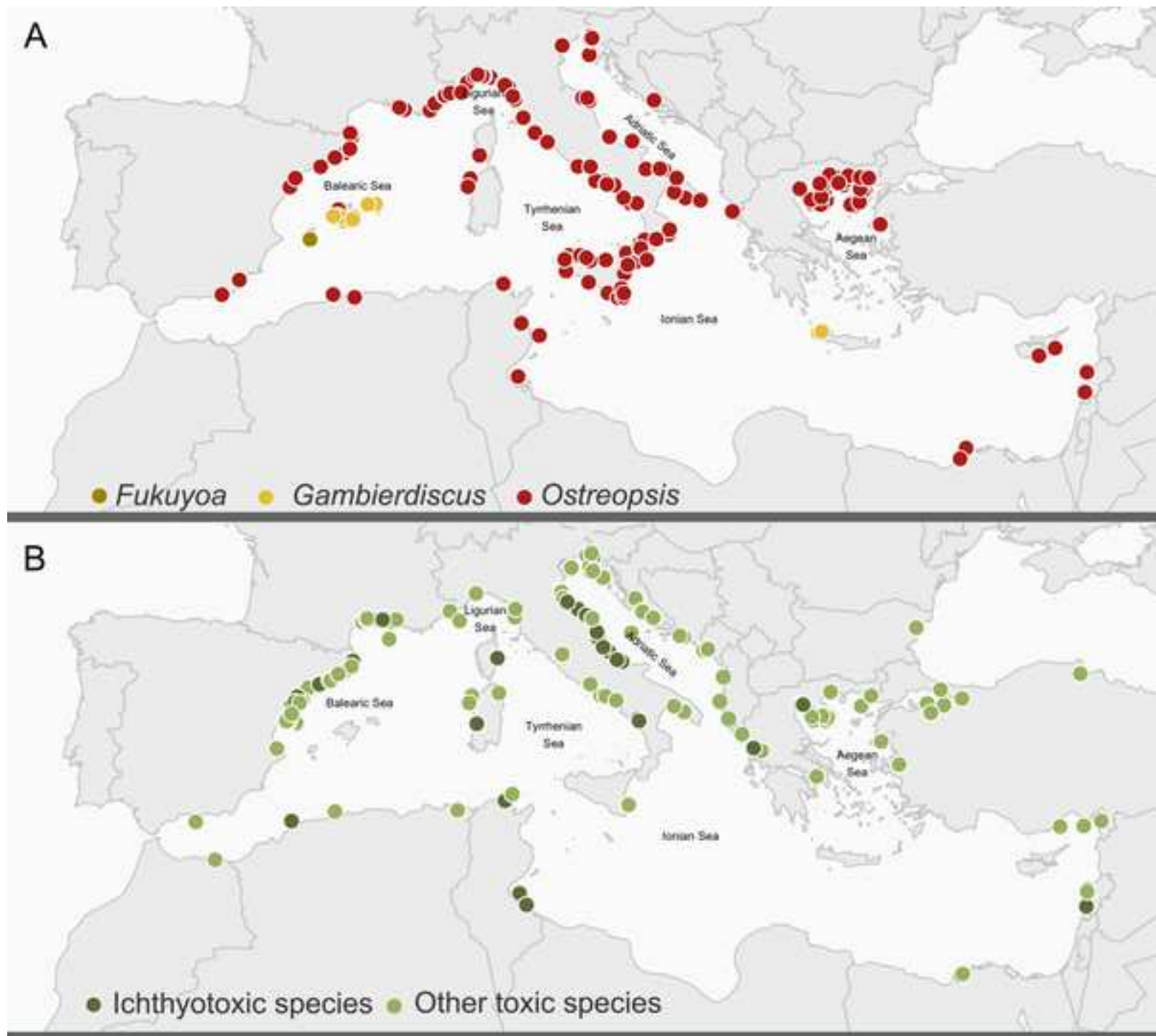
1529 not considered. A) Relative abundance of different types of nuisance with details of

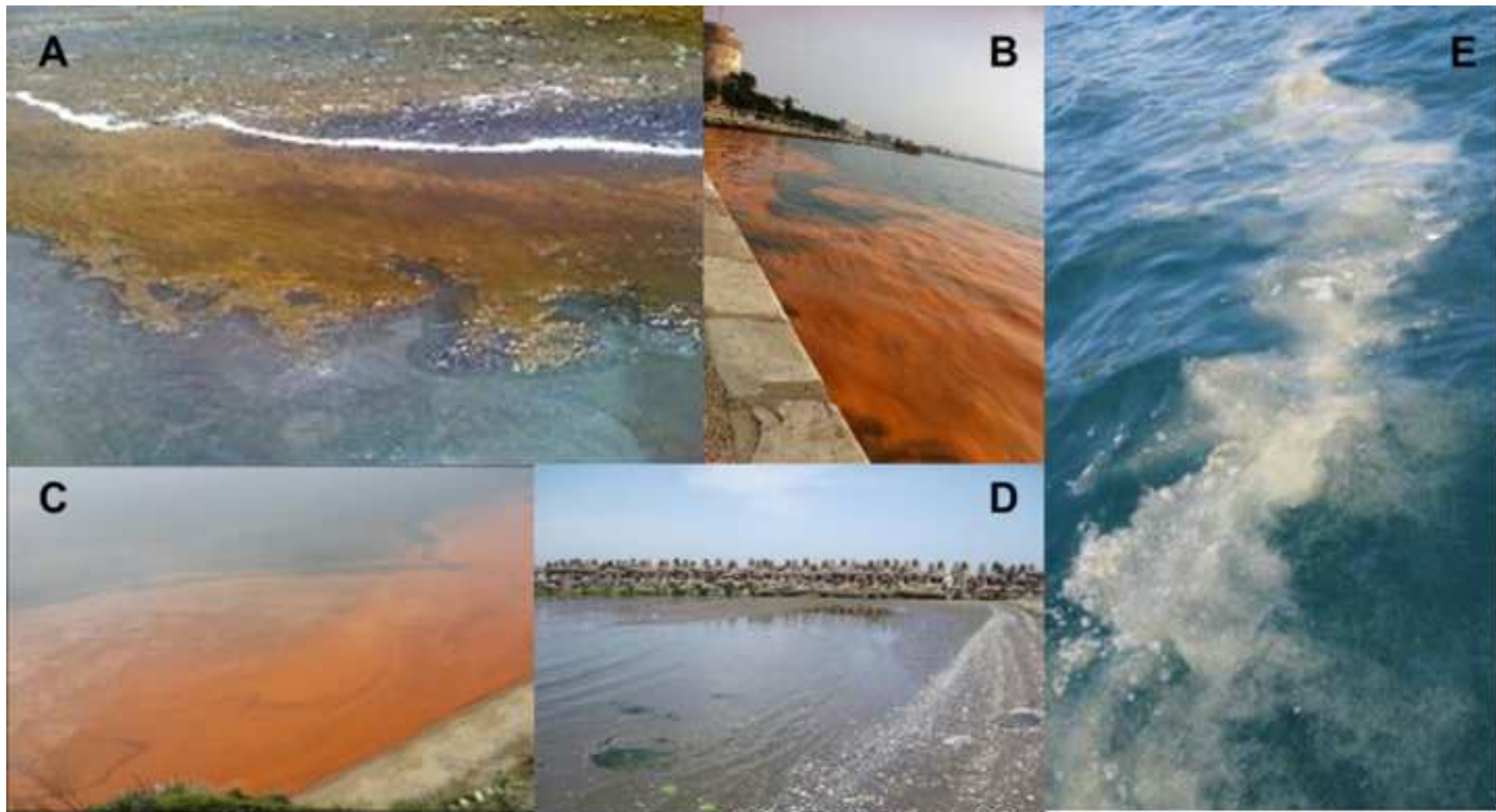
1530 seafood toxicity. B) Interannual variations of ASP, DSP and PSP toxicity events.

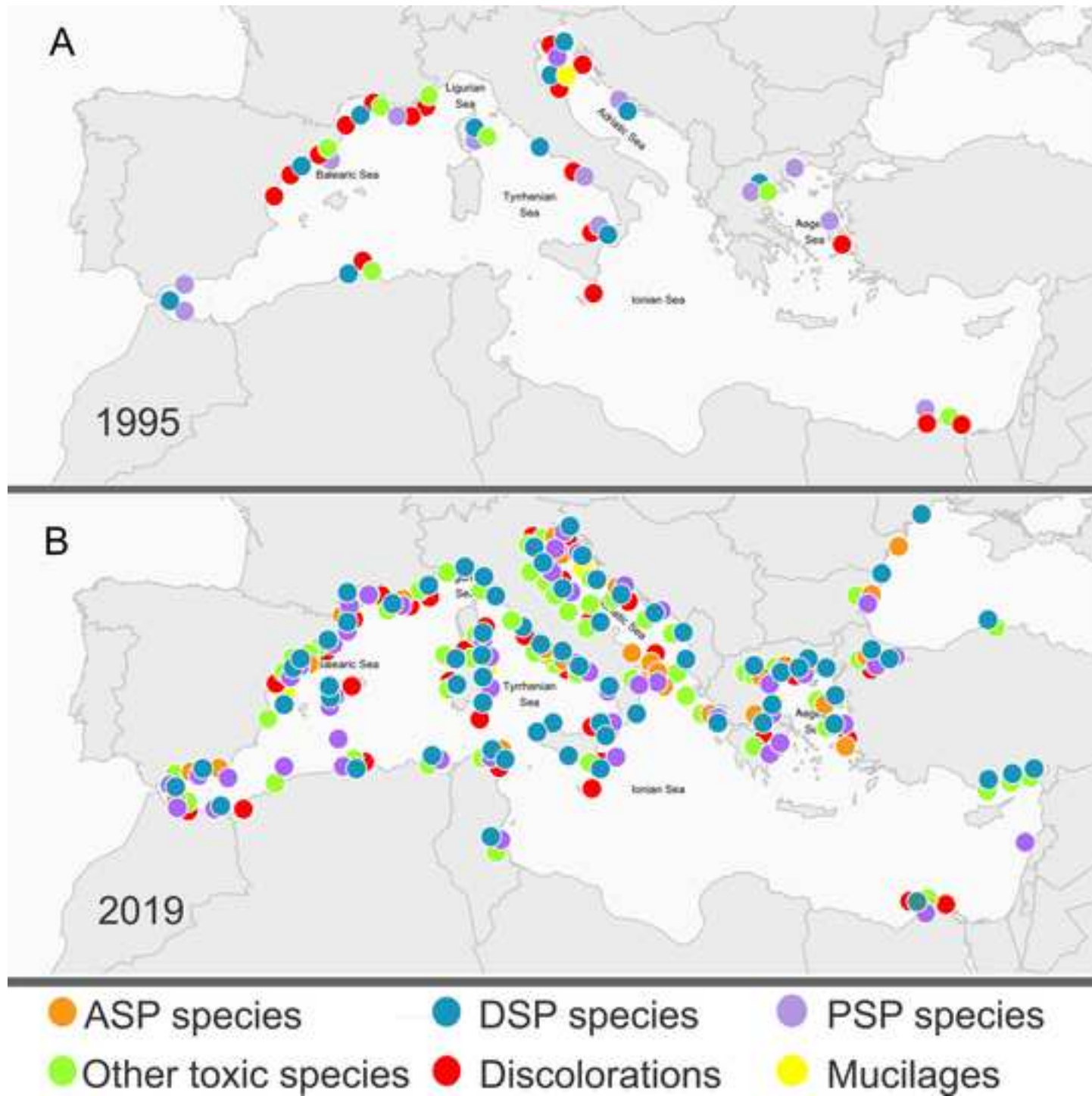












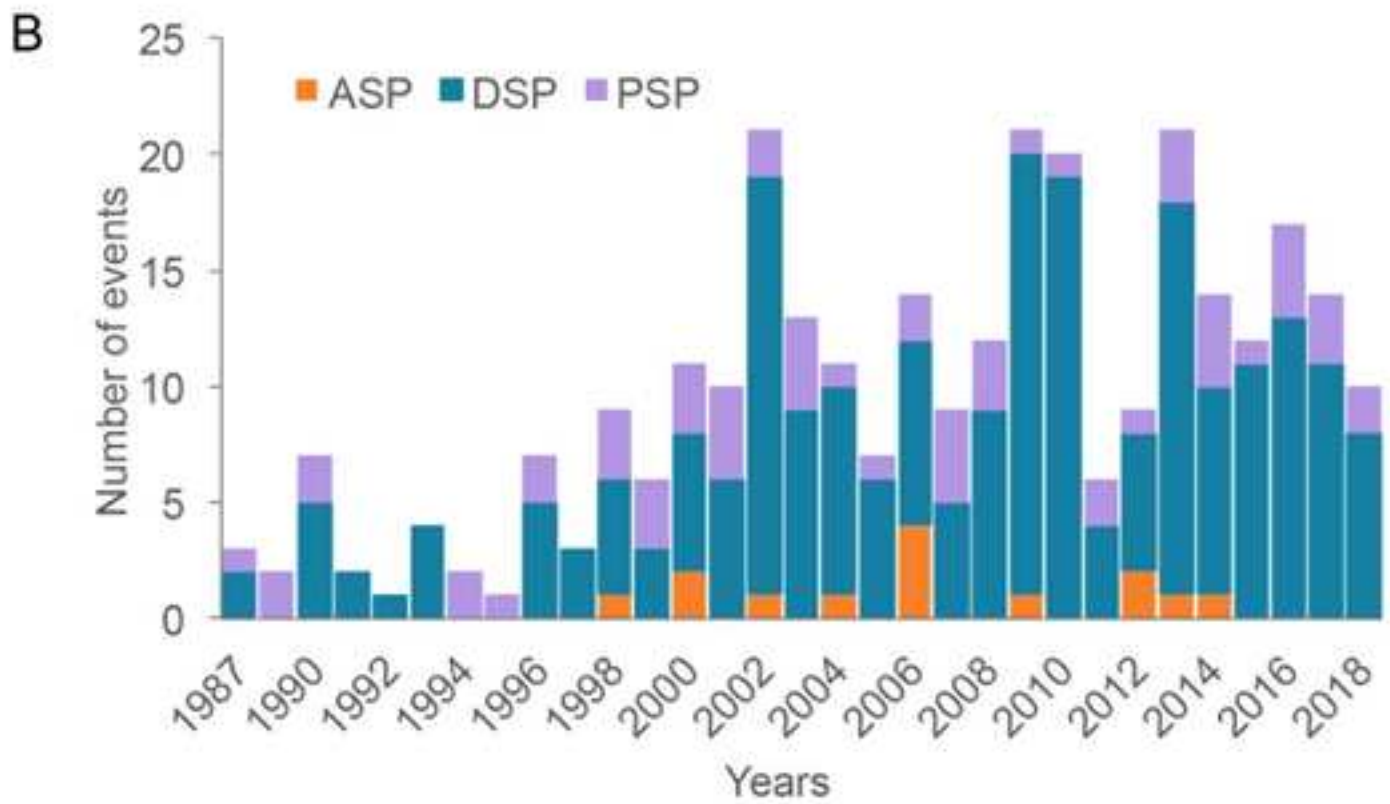
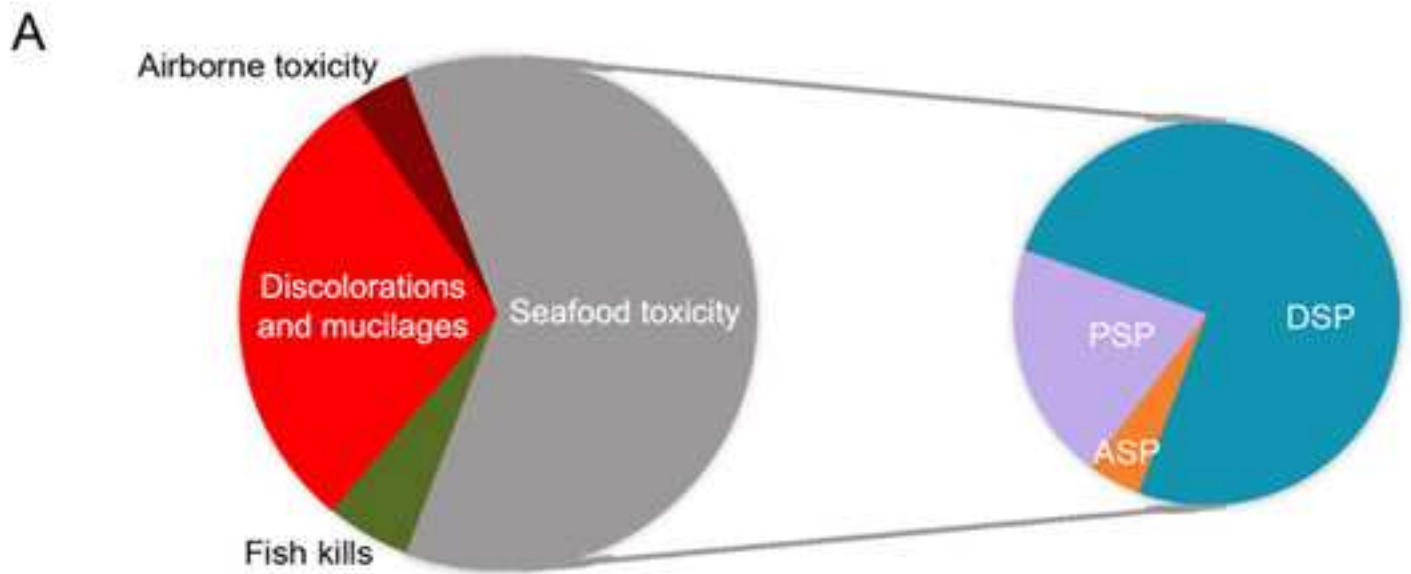


Table 1: Potentially toxic species in the Mediterranean Sea and associated types of syndromes or impacts (see Moestrup et al., 2009 and Lassus et al., 2016 for details). ASP, Amnesic Shellfish Poisoning; AZP, Azaspiracid Shellfish Poisoning; DSP, Diarrhetic Shellfish Poisoning; PSP, Paralytic Shellfish Poisoning; CFP, Ciguatera Fish Poisoning. ‘Other toxins’ include unknown toxins or toxins with poorly known effects.

Bacillariophyceae

<i>Halamphora coffeaeformis</i>	ASP
<i>Nitzschia bizertensis</i>	ASP
<i>Pseudo-nitzschia australis</i>	ASP
<i>Pseudo-nitzschia brasiliiana</i>	ASP
<i>Pseudo-nitzschia caciantha</i>	ASP
<i>Pseudo-nitzschia calliantha</i>	ASP
<i>Pseudo-nitzschia cuspidata</i>	ASP
<i>Pseudo-nitzschia delicatissima</i>	ASP
<i>Pseudo-nitzschia fraudulentata</i>	ASP
<i>Pseudo-nitzschia galaxiae</i>	ASP
<i>Pseudo-nitzschia hasleana</i>	ASP
<i>Pseudo-nitzschia multiseriata</i>	ASP
<i>Pseudo-nitzschia multistriata</i>	ASP
<i>Pseudo-nitzschia pseudodelicatissima</i>	ASP
<i>Pseudo-nitzschia pungens</i> ⁽¹⁾	ASP
<i>Pseudo-nitzschia subfraudulenta</i>	ASP
<i>Pseudo-nitzschia subpacificata</i>	ASP

Dictyochophyceae

<i>Pseudochattonella farcimen</i>	Ichthyotoxicity
<i>Pseudochattonella verruculosa</i>	Ichthyotoxicity
<i>Vicicitus globosus</i>	Ichthyotoxicity

Dinophyceae

<i>Alexandrium andersonii</i>	PSP
<i>Alexandrium balechii</i>	Ichthyotoxicity
<i>Alexandrium minutum</i>	PSP
<i>Alexandrium ostenfeldii</i>	PSP
<i>Alexandrium pacificum</i> ⁽²⁾	PSP
<i>Alexandrium pseudogonyaulax</i>	Ichthyotoxicity
<i>Alexandrium tamarense</i> ⁽²⁾	PSP
<i>Alexandrium taylorii</i>	PSP
<i>Amphidinium carterae</i>	Ichthyotoxicity
<i>Amphidinium klebsii</i>	Ichthyotoxicity
<i>Azadinium dexteroporum</i>	AZP
<i>Azadinium poporum</i>	AZP
<i>Dinophysis acuminata</i>	DSP
<i>Dinophysis acuta</i>	DSP
<i>Dinophysis caudata</i>	DSP
<i>Dinophysis fortii</i>	DSP
<i>Dinophysis infundibulum</i>	DSP
<i>Dinophysis ovum</i>	DSP
<i>Dinophysis sacculus</i>	DSP
<i>Dinophysis tripos</i>	DSP

<i>Fukuyoa paulensis</i>	CFP
<i>Gambierdiscus australes</i>	CFP
<i>Gambierdiscus belizeanus</i>	CFP
<i>Gambierdiscus carolinianus</i>	CFP
<i>Gambierdiscus silvae</i>	CFP
<i>Gonyaulax spinifera</i>	Other toxins
<i>Gymnodinium catenatum</i>	PSP
<i>Karenia bicuneiformis</i>	Ichthyotoxicity
<i>Karenia brevis</i>	Ichthyotoxicity
<i>Karenia cristata</i>	Ichthyotoxicity
<i>Karenia longicanalis</i>	Ichthyotoxicity
<i>Karenia papilionacea</i>	Ichthyotoxicity
<i>Karenia selliformis</i>	Ichthyotoxicity
<i>Karlodinium armiger</i>	Ichthyotoxicity
<i>Karlodinium corsicum</i>	Ichthyotoxicity
<i>Karlodinium veneficum</i>	Ichthyotoxicity
<i>Lingulodinium polyedra</i>	Other toxins
<i>Margalefidinium polykrikoides</i>	Ichthyotoxicity
<i>Ostreopsis fattorussoi</i>	Airborne disease
<i>Ostreopsis cf. ovata</i>	Airborne disease
<i>Ostreopsis cf. siamensis</i>	Airborne disease
<i>Pfiesteria piscicida</i>	Ichthyotoxicity
<i>Phalacroma mitra</i>	DSP
<i>Phalacroma rotundatum</i>	DSP
<i>Polykrikos hartmannii</i>	Other toxins
<i>Prorocentrum borbonicum</i>	Other toxins
<i>Prorocentrum cordatum</i>	Other toxins
<i>Prorocentrum emarginatum</i>	Other toxins
<i>Prorocentrum lima</i>	DSP
<i>Prorocentrum mexicanum</i>	Other toxins?
<i>Prorocentrum rhathymum</i>	DSP
<i>Protoceratium reticulatum</i>	Other toxins
<i>Vulcanodinium rugosum</i>	Other toxins

Haptophyceae

<i>Chrysochromulina leadbeateri</i>	Ichthyotoxicity
<i>Phaeocystis globosa</i>	Other toxins
<i>Prymnesium calathiferum</i>	Ichthyotoxicity
<i>Prymnesium faveolatum</i>	Ichthyotoxicity
<i>Prymnesium parvum</i>	Ichthyotoxicity
<i>Prymnesium polylepis</i>	Ichthyotoxicity

Raphidophyceae

<i>Chattonella marina</i> ⁽³⁾	Ichthyotoxicity
<i>Chattonella subsalsa</i>	Ichthyotoxicity
<i>Heterosigma akashiwo</i>	Ichthyotoxicity
<i>Fibrocapsa japonica</i>	Ichthyotoxicity

⁽¹⁾ Including *P. pungens* var. *aveirensis*

⁽²⁾ *A. pacificum* (group IV) and *A. tamarense* (group III), following the ribotype group designation in John et al. (2014) and Litaker et al. (2018)

⁽³⁾ Including *Chattonella marina* var. *antiqua*

Table 2: Potentially toxic species described from the Mediterranean Sea

Species name	Described in	Described as	Type locality
<i>Alexandrium minutum</i> Halim	Halim (1960)		Harbour of Alexandria, Egypt
<i>Alexandrium pseudogonyaulax</i> (Biecheler)	Biecheler (1952)	<i>Goniodoma</i>	Thau Lagoon, Gulf of Lion, France
Horiguchi ex K.Yuki & Y.Fukuyo		<i>pseudogonyaulax</i>	
<i>Azadinium dexteroporum</i> Percopo & Zingone	Percopo et al. (2013)		Gulf of Naples, Italy
<i>Chattonella subsalsa</i> Biecheler*	Biecheler (1936)		Saltern of Villeroy, Sète, France
<i>Dinophysis caudata</i> Kent	Kent (1881)		Nearby Fano, Marche Region, Italy
<i>Dinophysis fortii</i> Pavill.	Pavillard (1923)		Thau Lagoon and/or Sète harbour, France
<i>Dinophysis infundibulum</i> J.Schiller	Schiller (1928)		Southern Adriatic Sea
<i>Dinophysis sacculus</i> F.Stein	Stein (1883)		Kvarner Gulf, Croatia
<i>Dinophysis tripos</i> Gourret	Gourret (1883)		South of Ratonneau, Gulf of Marseille, France
<i>Karlodinium armiger</i> Bergholtz, Daugbjerg & Moestrup	Bergholtz et al. (2006)		Alfacs Bay, Catalonia, Spain
<i>Karlodinium corsicum</i> (Paulmier, Berland, Billard & Nézan) Siano & Zingone	Paulmier et al. (1995)	<i>Gymnodinium corsicum</i>	Diana Lagoon, Corse, France
<i>Nitzschia bizertensis</i> Bouchouicha-Smida, Lundholm, Hlaili & Mabrouk	Bouchouicha-Smida et al. (2014)		Bizerte Lagoon, Tunisia
<i>Ostreopsis fattorussoi</i> Accoroni, Romagnoli & Totti	Accoroni et al. (2016)		Batroun, Lebanon
<i>Prorocentrum lima</i> (Ehrenb.) F.Stein	Ehrenberg (1860)	<i>Cryptotomonas lima</i>	Sorrento, Gulf of Naples, Italy
<i>Prymnesium faveolatum</i> Fresnel	Fresnel et al. (2001)		Beach of Roquebrun, Cap Martin, France
<i>Vulcanodinium rugosum</i> Nézan & Chomérat**	Nézan and Chomérat (2011)		Ingril Lagoon, France

*A second, distinct genotype also discovered in Mediterranean waters (Klöpffer et al., 2013).

** First report in Rhodes et al. (2010) from New Zealand.