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(Article begins on next page)

1 **Innovative multifunctional finish for the improvement of Indoor Air Quality: performance at**
2 **laboratory and pilot scale**

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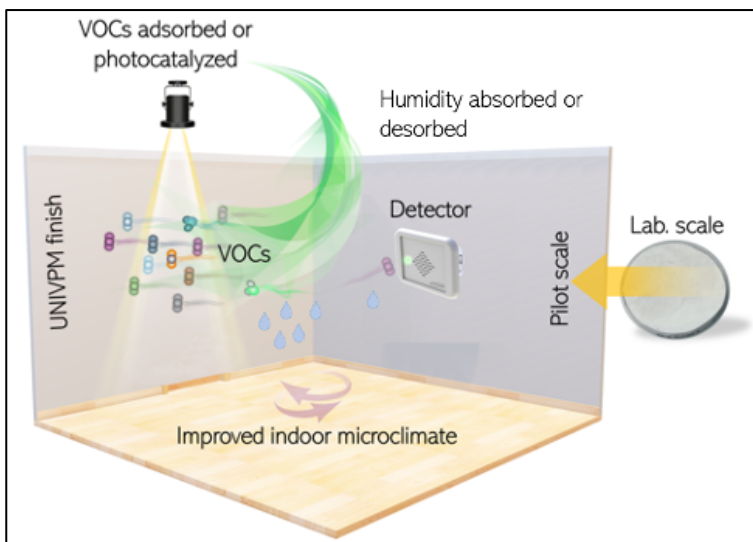
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12

13 **Graphical Abstract**



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15

16 **Abstract**

17 In this study, the impact of an innovative multifunctional finish on the Indoor Air Quality (IAQ) in terms of
18 temperature (T), relative humidity (RH), and volatile organic compounds (VOCs) was compared to that of a
19 commercial finish for the same application as reference. The two finishes were applied on the same commercial
20 substrate and tested both at laboratory and pilot scale. At laboratory scale, the water vapour permeability, the
21 moisture buffering capacity and the depolluting activity by adsorption of the innovative finish was 15%, 100%
22 and more than 200% higher than those of the commercial one on the same substrate, respectively. At pilot
23 scale, the innovative finish was more effective than the commercial finish in buffering humidity variations
24 (RH increases 53% slower compared to the commercial solution), temperature variations (-1.5°C), and
25 decreasing VOCs concentration (from 14% up to 63%, depending to the test conditions). Therefore, the
26 innovative multifunctional indoor finish has proven to be more effective than the commercial one in improving
27 IAQ not only in laboratory but also at pilot scale.

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Keywords

Indoor Finish, Indoor Air Quality (IAQ), De-polluting Properties, Moisture Buffering Capacity, Thermo-Hygrometric Properties, Pilot Scale Application.

Highlights:

- An innovative indoor finish was compared with a commercial one at lab and pilot scales
- The innovative finish was more effective to buffer humidity variations at lab and pilot scales
- The innovative finish had more depolluting abilities at lab and pilot scales
- The innovative finish is more effective in improving IAQ at lab and pilot scales

1. Introduction

Due to lifestyle changes people spend up to 90% of their daytime indoors in confined environments, such as houses, schools, offices, shopping centres, gyms, etc. [1]. Since the 2030 Agenda for Sustainable Development highlighted that all productive sectors should contribute to the reduction of greenhouse gasses (GHGs) emissions [2] and 40% of the global CO₂ equivalent emissions are produced by the construction sector [3] the current tendency is to provide more tightly sealed construction to create a stable internal environment that retains heat, thus leading to energy conservation.

However, the 13th goal of the 2030 Agenda “Climate Action” includes green and Nearly Zero Energy Buildings (NZEB) able not only to save energy and water, and reduce CO₂ emissions, but also to improve occupants’ health and well-being. But passive and near-zero energy projects use building materials with high thermal inertia and limit the implementation of Heating Ventilation Air Conditioning (HVAC) systems. Consequently, the air circulation needed to maintain a good indoor air quality (IAQ) for a healthy environment may be impeded which can affect the comfort of occupants. The increase of Volatile Organic Compounds (VOCs), moulds, airborne particles and inadequate relative humidity (RH) levels [4] [5] affects the comfort and health of occupants with both short- and long-term effects such as the well-known Sick Building Syndrome (SBS), a disease caused by poor-quality ventilation, recognized by the US National Institute for Occupational Safety and Health (NIOSH) [6] .

The need for more comfortable and healthier environments in combination with efficient buildings requires new approaches such as the development of passive systems that can assist the active ones, such as air conditioners, purifiers, or dehumidifiers, currently used to improve IAQ [7] [8]. In this context, finishes covering the surface of internal walls and ceilings, thanks to their high surface area exposed to indoor environment, can play an important role in passively improving IAQ [9] [10] [11] [12]. In finishes for indoor applications, highly porous materials have been demonstrated able to act both as RH buffers and as depolluting system to decrease VOCs concentrations. In this regard, it has already been confirmed that moisture transport, storage, and pollutants removal (for example toluene) is influenced by porous building materials [13] [14].

65 Concerning the depollution abilities of materials, adsorption is the most effective process whereby a gas or a
66 liquid (adsorbate) is passively removed from an air stream/atmosphere or fluid and transferred to the solid
67 surface of an adsorbent [15]. Consequently, to achieve high efficiency, adsorbent materials should have a large
68 specific surface area (approximately 300-3000 m²/g) to assure a high solid-fluid contact area [16]. In
69 depolluting processes, adsorption is often combined with another superficial phenomenon known as
70 Heterogeneous Photocatalytic Oxidation (PCO) [17] [18] [19] [20]. In PCO, under radiation (mainly UV),
71 reactive radicals from ambient O₂ and H₂O react with VOCs due to activated catalytic semiconductors such as
72 titanium dioxide (TiO₂). These additives can be incorporated either in bulk or applied as a coating on the
73 finishes to mineralize the adsorbed pollutants into less harmful substances [21] [22]. Hydraulic lime (HL)-
74 based products generally have higher depolluting properties than cement-based ones [23] due to the higher
75 presence of micro-pores and fewer nano-pores (gel pores) in the matrix which may hide the catalytic sites [24]
76 [25] [26] [27]. Indeed, in case of bulk addition of TiO₂, the PCO activities are significantly affected by
77 macro/micro pores and components of the finish [21] [22]. HL is also a more sustainable binder than cement
78 since its production has a lower energy impact [28]. Moreover, free-lime (CaO) which is part of HL
79 composition, can capture CO₂ (associated with greenhouse gases) by carbonation during the setting and
80 hardening of the binder, thus contributing to a more favorable CO₂ balance [29]. Sustainability and reduction
81 of indirect CO₂ can also be guaranteed including wastes or by-products in the binder or as aggregates [30] [31]
82 [32] [33].

83 Concerning indoor humidity control by materials, indoor finishes with high water vapour permeability and
84 Moisture Buffering Capacity (MBC) [34] can guarantee the indoor optimal RH ranges [35] of about 40% [36].
85 Earth plasters have a greater effect on the hygrothermal balance if compared to air lime, gypsum, and cement
86 plasters [37]. Many laboratory-scale studies aimed at improving IAQ by means of adsorption/PCO in paints,
87 plasters and finishes have been reported in the literature [38], including real scale experimental analysis [39]
88 [40]. Vega-Mendoza et al. tested the self-cleaning and CO₂ remediation of an alkali-activated fly ash
89 photocatalytic mortar coating incorporating α/β -Bi₂O₃ homojunctions as a photocatalyst with high visible-light
90 response [41] in a batch reactor. Jeon et al. focused on the improvement of IAQ by using a toluene adsorption
91 paint with activated carbon in an inorganic binder both in laboratory static conditions and in an urban
92 residential apartment [42]. A photocatalytic TiO₂-based paint was also tested to decrease NO₂ concentrations
93 in a lecture room of the Technical University of Madrid [43].

94 Some of the present authors have proposed an innovative inorganic finish (Patent IT 102017000033750 –
95 Multifunctional Mortar) which has been demonstrated to improve IAQ in terms of MBC and depolluting ability
96 at laboratory scale [44] [45] [46]. In the patented mix, depolluting abilities are enhanced by coupling
97 unconventional adsorbent aggregates to photocatalytic TiO₂ particles. The use of adsorbents currently not
98 employed in construction, but in chromatographic techniques and filters for water and air purification, allows
99 to increase lightness, permeability to water vapour and buffer properties of the finish. HL and non-hazardous
100 special wastes, such as fly and bottom biomass ashes, not only ensure good mechanical properties but also the
101 environmental sustainability of the product. However, it is well known that the results obtained at the

102 laboratory scale may be overestimated [47] [48] and have no practical relevance when moving to real scale.
103 Also, the influence on thermo-hygrometric properties of the whole wall package [49] is of utmost importance
104 [50] , since finishes are applied on indoor surfaces.

105 Recent literature shows a limited number of studies that demonstrate passive VOCs removal by mortars relying
106 on the combined exploitation of adsorbing and photocatalytic properties of the innermost surface of the
107 building envelope. In fact, generally literature studies have been carried out at laboratory scale and for outdoor
108 applications; in some cases, the reproduction of the real environment was done through modeling [51].
109 Moreover, the depolluting materials tested in real environments [38] are commonly incorporated into building
110 components (i.e. ceilings, flooring, plasters or plasterboards) rather than in finishing materials. Finally, to the
111 best of the authors' knowledge, there are no previous works based on the comparison of multiple properties
112 such as thermo-hygrometric and depolluting ones in on-site applications.

113 Therefore, the aim of this paper is to investigate the properties of the multifunctional innovative finish mix
114 compared to a commercial one, both at laboratory and pilot scales. Both the patented and commercial finishes
115 were applied on the same commercial substrate and compared in terms of thermo-hygrometric and depolluting
116 properties in lab scale tests. Then in pilot scale tests, the patented finish and the commercial reference were
117 applied on the same commercial substrate in two different dedicated test boxes under the same environmental
118 conditions. The indoor parameters, such as Temperature (T), VOCs concentration and RH, of the boxes were
119 compared to verify the benefits given by the multifunctional finish to IAQ.

120

121 2. *Materials*

122 The innovative finish (labelled “U” in the following text) was prepared with a commercial natural hydraulic
123 lime (NHL), a mix of adsorbent materials and special wastes for aggregates, TiO₂ filler and commercial
124 admixtures.

125 The commercial NHL was the product i.pro CALIX BLANCA NHL 3,5 (Heidelberg Materials – Italcementi)
126 according to the standard UNI EN 459-1, with a declared density of 2700-3000 kg/m³.

127 The selected aggregates were a mix of adsorbent materials not commonly used in the construction sector, and
128 fly/bottom ashes from the thermal treatment of biomasses. The adsorbent aggregate had a water absorption
129 capacity to reach the saturated surface dry (ssd) condition of 86%, a density of 1310 kg/m³ and a maximum
130 grain size diameter (D_{max}) of 300 μ m. The biomass ashes came from an Italian biomass-to-energy valorisation
131 plant, located in Rieti (Lazio region). The water absorption capacity was 49% and 20%, the ssd density was
132 1410 and 1940 kg/m³ and D_{max} was 1000 and 210 μ m, for the bottom and fly ashes, respectively [52].

133 Aeroxide® by Evonik, with nano-size particles of 20–50 nm, was chosen as TiO₂ photocatalytic addition. The
134 specific surface area was 52 m²/g, the micropore area 0.36 m²/g and the average pore diameter (4V/A) 146 Å,
135 as measured by BET (Gemini 2360 Surface Area Analyzer equipped with FlowPrep 060, micrometrics). The
136 pH of a 4% water dispersion was 3.5–4.5, and the density was 3.1 g/cm³ as declared in the technical datasheet.
137 The amount of TiO₂ added to the finish was 2% of total powder mass.

138 To optimize the fresh properties, facilitate the application, and avoid shrinkage cracking after the application
139 on the substrate, 3 commercial admixtures were added to the patented mix when applied at pilot scale: i)
140 shrinkage reducing admixtures (SRA) at 0.5% of total powder mass; ii) micro cellulose at 0.7% of total powder
141 mass and iii) superplasticizer at 0.02% of total powder mass.

142 The U finish was prepared by adding water until a proper workability for the application was reached
143 (measured as indicated in section 3.2.), which was about 40% of the total weight of powders. It was applied
144 on the commercial substrate in two layers for a total thickness of 3 ± 0.5 mm.

145 The properties of the U finish were compared to those of a commercial one, limepaint by Diasen® (labelled
146 “L” in the following text), used as a reference. L is a water-based paint currently used as a ready-to-use finish
147 for lime-based materials that can be applied on walls, renders, and ceilings.

148 Calce Storica by Diasen® (C), was selected as commercial substrate for the application of both L and U finish.
149 C is a premixed mortar composed by commercial aggregates and natural hydraulic lime as binder, prepared by
150 adding 18% of water, as indicated in the technical datasheet. The two finishes were applied after 15 days of
151 curing: at lab scale, curing was performed at T of 20 ± 2 °C and RH of $60 \pm 5\%$, while at pilot scale at room
152 conditions with T ranging from 18 to 20 ± 2 °C and RH from 80 to $60 \pm 5\%$.

153 For on-site applications, the commercial substrate C was applied for 20 mm on the entire wall for 2 cm. Then,
154 the U finish was applied in 2 consecutive layers, approximately 1.5 mm thick each, and TiO₂ was added only
155 in the finish of the outermost layer. Following the technical datasheet, a layer of fixative (D20 supplied by
156 Diasen®) was applied before the L finish. The labels and descriptions of all tested materials are reported in
157 Table 1.

158

159 Table 1 label and description of materials

Label	Description
U	Innovative finish: a hydraulic lime-based finish with adsorbent materials, industrial by-products, TiO ₂ filler and commercial admixtures
L	Commercial finish: ready-to-use lime-based water paint
C	Commercial base coat: premixed hydraulic lime-based mortar
C-U	Finish U applied on base coat C
C-L	Finish L applied on base coat C

160

161 3. Laboratory tests methods

162 The full characterization of the innovative multifunctional finish (U) in terms of mechanical, thermo-
163 hygrometric, mould growth inhibition and depolluting properties, has been recently published [44].

164 The **mineralogical composition** of the ashes was investigated by X-ray analysis, using a RX Philips PW 1730
165 X-ray diffractometer (operating voltage/current 40 kV/30 mA; scan mode: continuous speed, 3°/min). The
166 main components were quantified through TGA (Mettler Toledo) at speed rate of 10 °C/min under oxidizing
167 atmosphere.

168 The **workability** was evaluated by slump test according to the UNI EN 1015-3:2007 standard and resulted in a
169 stiff consistency, as classified in UNI EN 1015-6:2007.

170 **The compressive strength** test was carried out according to UNI EN 1015-11:2007 on three $4 \times 4 \times 16$ cm
171 specimens by means of a hydraulic press (Galbaldini - Cardano al Campo (VA) Italy) with a precision of 1%
172 and the average results have been reported.

173 The **thermo-hygrometric properties** were evaluated on three specimens for each mix, and the average values
174 reported.

175 The thermal conductivity (λ) of C-L and C-UNI was measured at $T = 20 \pm 2$ °C and $RH = 50 \pm 5\%$ according
176 to UNI EN 12664:2002, following equation (1):

$$177 \lambda = \frac{Q \cdot s}{A \cdot (T_2 - T_1)} \quad (1)$$

178 where Q/A is the heat flux (W/m^2), s is the sample thickness (m), and T_1 and T_2 (K) are the temperatures at the
179 two different sides of the sample, set at 15 °C and 30 °C.

180 The hygrometric properties were evaluated both in static and dynamic conditions, in terms of water vapour
181 resistance factor (μ) and Moisture Buffering Value (MBV), respectively.

182 The water vapour permeability was tested according to the UNI EN 1015-19:2007, and data were processed
183 according to the UNI EN ISO 12572:2007 standard, which evaluates the μ value of cylindrical specimens ($h =$
184 3 cm, $d = 20$ cm). The tested specimens were C-L, C-U and C.

185 The MBV defines the humidity absorbed and desorbed by the material on the surface unit when it is placed in
186 an environment at RH of 75% for 8 hours and of 33% for 16 hours, according to a simplified procedure of
187 NORDTEST, which evaluates the MBC of the material [53] [54] [55]. For each type of specimen (C, C-L, and
188 C-U), for each specimen the average MBV calculated over 3 cycles was obtained. The tested specimens were
189 cylindrical ($h = 3$ cm, $d = 10$ cm) with the lateral surface previously sealed with silicone to guarantee a
190 homogeneous water vapour flow.

191 The **depolluting properties** of the finishes were tested on cylindrical specimens ($h = 0.8$ cm, $d = 8$ cm) in terms
192 of adsorption and photocatalytic activity [46].

193 The adsorptive properties were evaluated by a batch test where Methyl Ethyl Ketone (MEK) was used as a
194 model pollutant. The specimen was placed in a glass box of 17 l, with a fan on the bottom, and 50 μ l of MEK
195 was injected with a micro-syringe from the upper hole hermetically closed with a rubber sept, to obtain inside
196 the box a theoretical initial MEK concentration of 797 ppm. The MEK concentration inside the box in time
197 was evaluated by collecting 10 μ l of air every 8 mins for 90 mins and analysing MEK concentration by a gas
198 chromatograph (Carlo Erba Gas 8000 Top; Flame Ionization Detector; injector split 1:15; carrier flow 2
199 ml/min; capillary column, 25 m \times 0.32 mm, 0.52 μ m cross-linked methyl siloxane; isotherm condition 40 °C).

200 The photocatalytic activity was assessed according to the UNI 11247:2009 standard, which measures the
201 nitrogen oxides (NO_x) degradation rate in plug flow condition. C-L and C-U were tested both under VIS
202 (Philips lamp, 42W) with intensity equal to 9 ± 1 W/m^2 as detected by Delta Ohm HD 2101.1 photo radiometer
203 (probe LP471 RAD 400 – 1050 nm), and UV radiation (halogen lamp, 400W) with measured light intensity
204 of 20 ± 1 W/m^2 (probe LP471 UVA 315 – 400 nm). The NO_x flux was supplied by a NO_x tank (SAPIO S.r.l.,

205 Monza, Italy), for a concentration of 500 ± 5 ppb in a dilution flow of 1.5 l/min, and was kept constant by
 206 mixing with air at $T = 25$ °C and $RH 50 \pm 10\%$ using a dilution system (Calibrator 8188, Rancon Instruments
 207 S.p.A., Milan, Italy) continuously monitored in terms of NO_x and NO concentrations by a chemiluminescence
 208 NO_x analyser (nitrogen oxide analyser model 8841; Monitor Labs, Englewood, CO, United States). Each
 209 tested specimen was placed in a 3 l borosilicate reactor and, as soon as the NO_x flux was stable, photocatalytic
 210 activity (%D) was evaluated following equation (2):

$$211 \quad \%D = \frac{C(\text{dark}) - C(\text{light})}{C(\text{dark})} 100\% \quad (2)$$

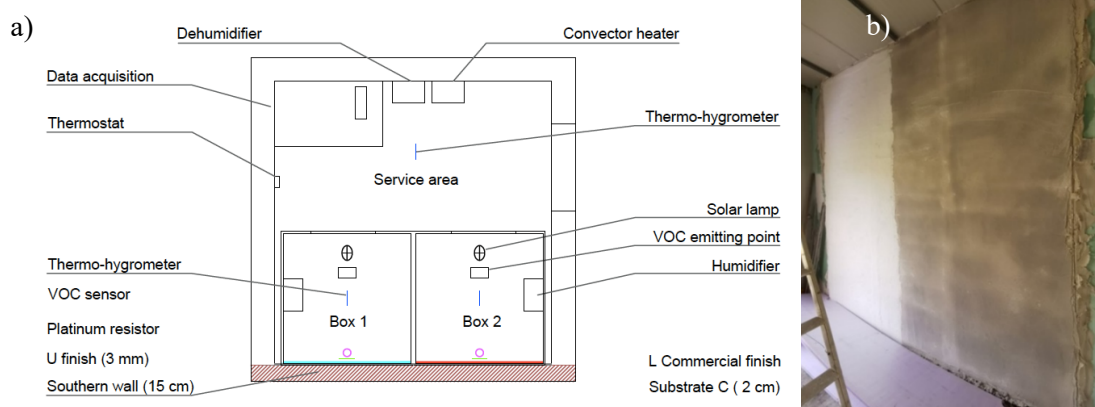
212 where %D is the percentage of degradation and C(dark) and C(light) are the concentrations of NO or NO_x in
 213 the dark and light conditions (after 10 min of irradiation), respectively.

214

215 4. On-site experimental methods

216 For on-site monitoring, a test room has been set up (Figure 1). The area identified for the experimentation is
 217 located in Ancona, central Italy (latitude: 43.54°, longitude: 13.37°, altitude: 137 m). The experiment was
 218 conducted in a prefabricated test room, measuring approximately $3 \times 3 \times 3$ m³ characterized by an aluminum
 219 sandwich panel envelope with an internal layer of mineral wool 12 cm thick.

220 On the north-eastern and western sides, a further 10 cm thick external insulation layer of polystyrene with an
 221 aluminum finish gave the walls a total thickness of 22 cm. On the southern side, the insulation layer thickness
 222 was 12 cm and, externally, a white rendered cement board was inserted, reaching a total thickness of 15 cm.
 223 The internal volume was divided in half: an antechamber hosting the computer for monitoring and a room on
 224 the southern exposure, which was in turn divided into two sub-rooms BOX1 and BOX2, (each with dimension
 225 of $1.2 \times 1.2 \times 2.9$ m³), hosting on the internal wall facing south the innovative C-U finish and the commercial
 226 C-L finish, respectively.



227

228 Figure 1. a) Plan of the experimental room and b) image of the two finishes before the test room separation

229

230 The two test boxes were closed with plasterboard panels and sealed with adhesive tape. A data acquisition and
 231 remote microclimatic control system was set up outside the two boxes, in the service area. The system provided
 232 the continuous data acquisition of: i) VOCs concentration inside box 1 and box 2 (VOC1, VOC2), ii) Air
 233 Temperature inside box 1 and box 2 (T1, T2), iii) Surface Temperature of the two finishes (RTD1, RTD2), iv)

234 Air Temperature of the service area (T_{ser} , v) Relative Humidity inside box 1 and box 2 (RH1, RH2) and
 235 Relative Humidity of the service area (RH_{ser}). Inside each box, the irradiance was measured before the
 236 beginning of the tests to assess the photocatalytic properties of the finish. Table 2 provides the information
 237 about the sensors.

238

239 Table 2. List of the sensors used in the experiment.

Parameter	Sensor type	Model	Range	Accuracy	Area of installation
Air temperature	Thermo-hygrometer	PCMini52, Michell	-20 °C to 80 °C	± 0.2 °C	Service area Test area
Surface temperature	Platinum resistor	PT100	-40 °C to 80 °C	± 0.05 °C	Test area
Relative humidity	Thermo-hygrometer	PCMini52, Michell	0 to 100%	<±2%	Service area Test area
VOCs concentration	Photoionization detector - PID	Series 900, Aeroqual	0 to 30 ppm	±0.02 ppm +10%	Test area
Irradiance	Photo-radiometer	HD 2102.2, DeltaOhm (Probes LP471UVA, LP471RAD)	0.1 mW/m ² to 2000 W/m ²	±1digit	Test area, only for the test of irradiance

240

241 In the service area, a control system based on LABVIEW allowed to maintain constant within the two boxes
 242 RH and T values = $55 \pm 5\%$, and 20 ± 0.5 °C, respectively.

243 Inside each box a humidifier (Medisana, model V-401671, 4.2 l capacity) and two different sources of radiation
 244 (lamps), namely a solar lamp (Ultravita LUX Osram, 300W) and a UV lamp (RepTech 100W), were installed.
 245 The measured irradiances were 3 and 8 W/m² in case of UV radiation and 18 and 30 W/m² in case of VIS
 246 radiation for the solar lamp and the UV lamp, respectively. When necessary, inside each box a MEK emitting
 247 point was placed to act as an internal source of VOCs.

248 To wait for proper curing of materials, the data acquisition of the environmental parameters of the two boxes
 249 started one month after the application of the finishes.

250 The following conditions were explored:

251 • **Temperature increase:** from 20 to 28 °C by switching on the UV lamp with UV = 3 W/m² and VIS = 18
 252 W/m². Tests were conducted in dark conditions and under UV/VIS radiation provided by the UV lamp, to
 253 evaluate the possible photocatalytic activity. The irradiance of UV/VIS measured by a photo-radiometer on
 254 the surface of the pilot wall, perpendicularly to the lamp, was equal to 8 W/m² and 30 W/m², respectively.

255 • **RH increase:** by switching on the humidifier.

256 • **VOCs concentration change:** by MEK (used as a model pollutant) static or spot load or by UV lamp
 257 switched on. For the static load test a beaker with 10 ml of MEK was placed in the center of each box floor;
 258 for the spot load test, 10 µl and 50 µl of MEK was injected inside each box by a graduated syringe.

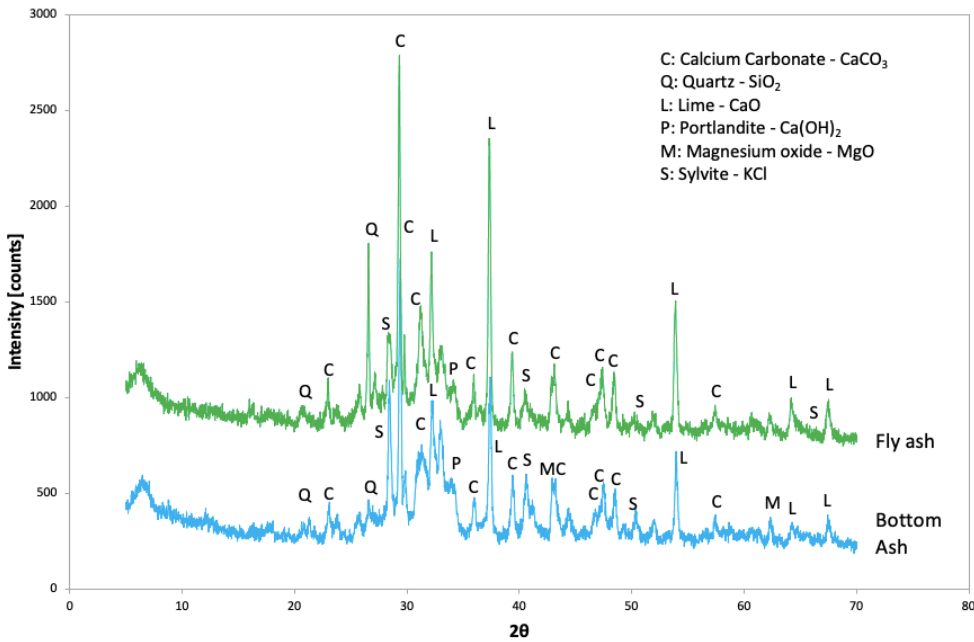
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260 5. Laboratory test results

261 5.1. Preliminary evaluations on ashes composition and physical-mechanical properties of the
262 innovative finish

263 The mineralogical composition of ashes was characterized by X-ray diffraction analysis (Figure 2). The main
264 detected compounds were calcium carbonate (CaCO_3), quartz (SiO_2), calcium oxide (CaO) and Portlandite
265 ($\text{Ca}(\text{OH})_2$); the last two calcium-based compounds can contribute to the binding phase. As crystalline species,
266 traces of magnesium oxide (MgO) and sylvite (KCl) were also detected but the low concentration of the latter
267 is confirmed by the absence of efflorescence on the surface of the prepared specimens.

268



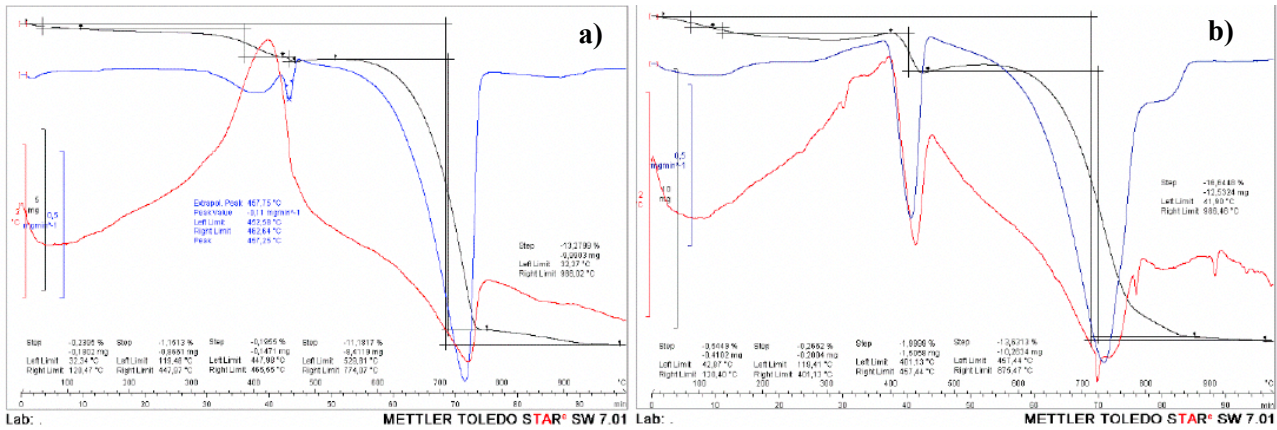
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270 Figure 2. XRD analysis of fly and bottom ashes.

271

272 TGA quantified the main components of the biomass ashes (Figure 3). The weight loss up to approximately
273 150 °C through an endothermic process is correlated to the loss of un-bound H_2O . The exothermic process
274 from approximately 120 to 450 °C is due to the combustion of organic material; the unburnt organic carbon
275 with a related exothermic reaction at T around 400-450 °C is detected only in fly ash. From 470 to 550 °C the
276 endothermic weight loss is due to the decomposition of $\text{Ca}(\text{OH})_2$ whereas the final endothermic weight loss
277 up to about 800 °C is due to the decomposition of CaCO_3 , quantified as 25.4% and 30.0% of the total content
278 for fly and bottom ashes, respectively. $\text{Ca}(\text{OH})_2$ is more present in fly ash than in bottom ash, with contents of
279 8.2% and 0.8%, respectively.

280



281
282 Figure 3. TG/DTA of fly (a) and bottom (b) ashes.

283
284 The workability of the U finish was evaluated through the slump test and resulted in a stiff consistency, as
285 classified in the UNI EN 1015-6:2007 standard.
286 The compressive strength of U finish was 8.7 MPa, a higher value compared to that generally obtained by a
287 3.5 NHL-based mortar, probably due to calcium-based compounds such as CaO and Ca(OH)₂ in the added
288 ashes which contribute to the binding phase.

289 **5.2 Thermo-hygrometric properties of studied mortars and systems**

290 The thermal conductivity values for C (substrate), C-L and C-U were: 0.112, 0.119 and 0.123 W/mK,
291 respectively. As expected, the presence of a finish, regardless if commercial or innovative, had a negligible
292 effect on λ value: indeed, the C-U system had less than 5% higher thermal conductivity than C-L and C.
293 The μ value was 21 for C-L and 18 for C-U demonstrating that the water vapour permeability is mainly affected
294 by the substrate C ($\mu = 19$).
295 The MBV values were 0.26, 0.15 and 0.34 g/(% m²) for C, C-L and C-U, respectively; this means an increase
296 of the hygroscopic buffer capacity of more than double when the innovative U finish is applied.

297 **5.3 Depolluting properties**

298 As regards depolluting properties, the results are presented in terms of Ci/C0 percent where C0 is the
299 theoretical initial concentration of MEK in the box and Ci is the concentration in the box during the air
300 sampling. After 90 mins of testing, C-L presented an adsorption capacity of 19% while C-U of 56%.
301 The photocatalytic activity under UV radiation was detected only in C-U specimen with TiO₂, and degradations
302 of NO and NO_x of 28% and 25% were measured, respectively. In case of C-L, no degradation was detected as
303 expected due to the absence of a photocatalytic agent [46]. The results are shown in Table 3.

304
305 Table 3. Summary of the results obtained by lab scale tests

Parameter		C	C-L	C-U
Thermal conductivity, λ	W/(mK)	0.112	0.119	0.123
Water vapour resistance factor, μ	-	19	21	18

Moisture Buffering Value (MBV)	g/(% m ²)	0.26	0.15	0.34
Adsorption capacity (MEK)	%	-	19	56
Photocatalytic degradation of NO	%	-	0	28
Photocatalytic degradation of NO _x	%	-	0	25

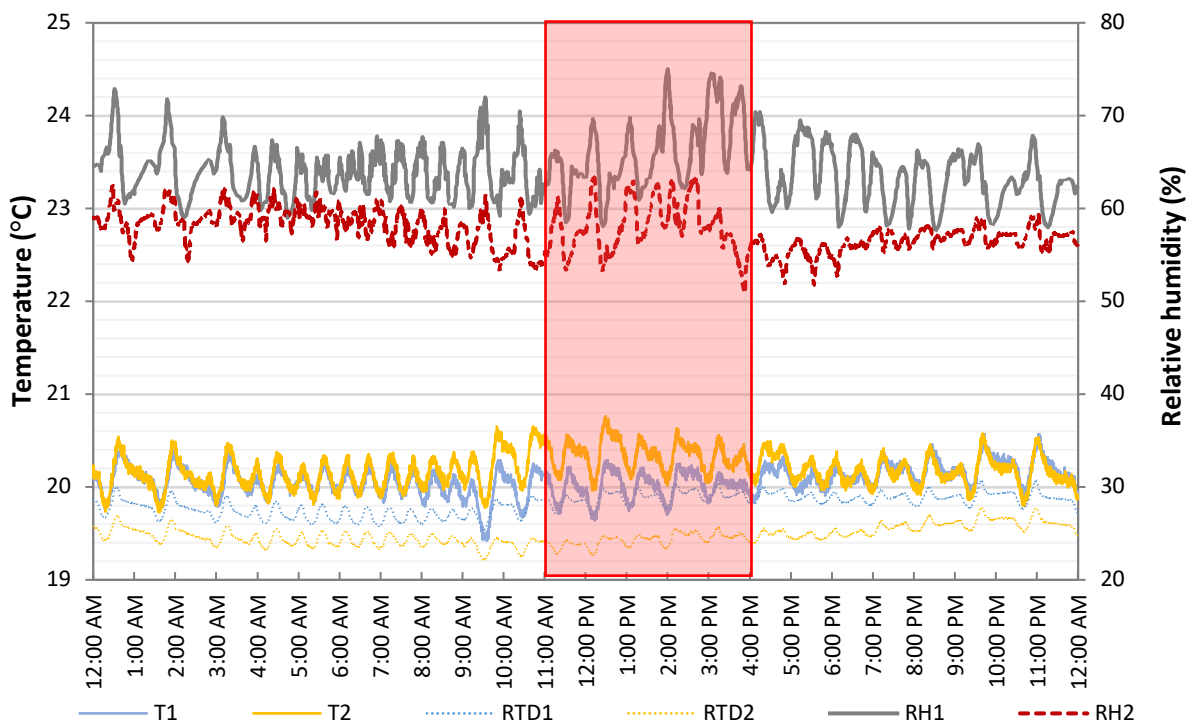
306

307 6. On-site tests results

308 6.1. Comparison under free running conditions

309 T and RH values recorded one month after the application of finishes in the two boxes are shown in Figure 4,
 310 the typical conditions with no external changing. The box with the U finish always showed approximately 5%
 311 higher RH (named RH1 in the figure) than traditional L finish (RH2), probably due to the higher water content
 312 of the U-finish. In the two boxes, the T values were similar (20 ± 1 °C) at night and during the non-central
 313 hours of the day. On the other hand, in the warmest hours of the day (from 11:00 AM to 4:00 PM), the
 314 innovative finish revealed slightly reduced T peaks with a more flattened trend and maximum achieved
 315 temperature levels around 0.5 °C lower than C-L, indicating a greater thermal inertia.

316



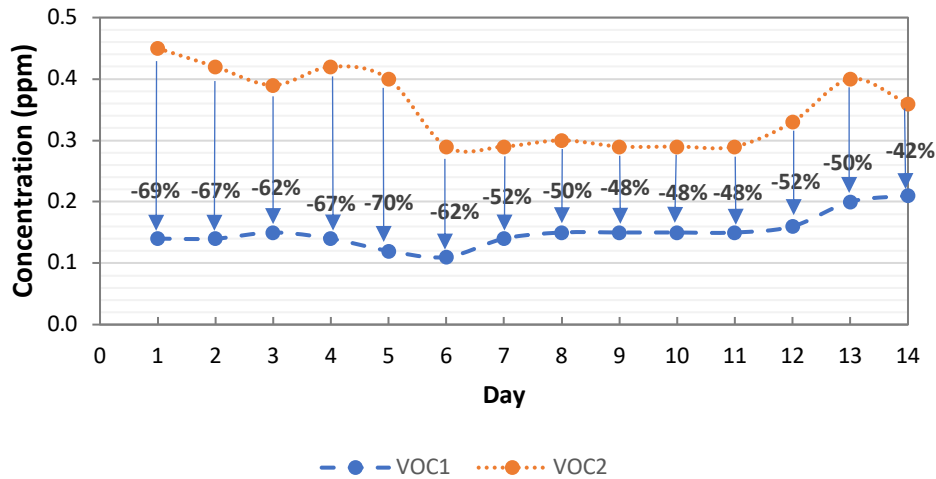
317

318 Figure 4. Initial conditions: 24 hours of monitoring of the two boxes (day 10 of monitoring). The warmest hours are
 319 highlighted. The number 1 is for box 1 (C-U) and number 2 for reference box 2 (C-L).

320

321 Figure 5 shows the average VOCs concentration recorded for two selected weeks of the monitored period.
 322 During the first week, up to 70% higher concentration of VOCs was observed in the box with the commercial

323 L finish compared to the one with the U finish. After this period, the VOCs concentration recorded in the box
 324 with C-L reached an average daily value from 0.40 ppm to 0.30 ppm, while in the box with C-U from 0.20
 325 ppm to 0.15 ppm. This can be attributed both to the lower VOCs emissions of the U finish compared to the L
 326 finish and to the higher VOCs adsorption capabilities of the U finish compared to the commercial L finish.



327
 328 Figure 5. Initial monitoring period (14 days): initial VOCs concentration detected in box1 and box 2. The number 1 is for
 329 box 1 (C-U) and number 2 for reference box 2 (C-L).
 330

331 **6.2. Comparison under high indoor temperatures**

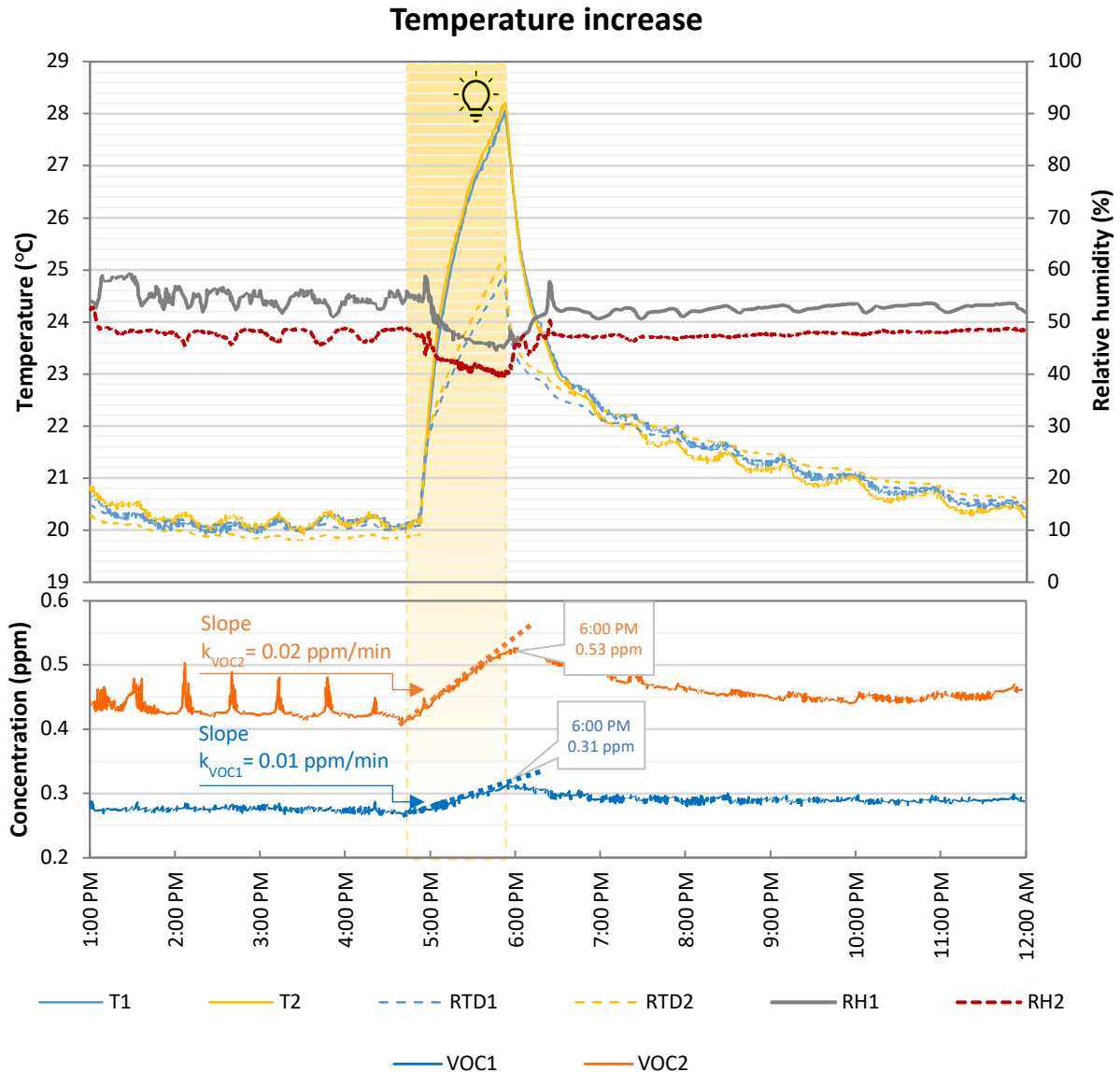
332 Figures shows the results obtained by increasing the indoor temperature from 20 °C to 28 °C, by turning on
 333 the solar lamp for 1 hour in each box to simulate warm days or household activities.

334 The rise of air temperatures in the two boxes (T1 and T2) was the same. However, 1 hour after the lamp turning
 335 off, T1 was slightly higher than T2 due to the higher thermal inertia provided by the U finish compared to the
 336 L finish. For the same reason, after switching on the lamps, the surface temperature of the U finish (RTD1)
 337 was approximately 0.5 °C lower than RTD2.

338 Figure 6 shows also the VOCs concentration variations with the temperature increase. In general, data showed
 339 an increase in VOCs released under higher temperature. After 1 hour of irradiation the VOCs concentration
 340 increased: inside box 1 with the new finish by 13% (from 0.27 ppm to 0.31 ppm); inside box 2 by 21% (from
 341 0.43 ppm to 0.53 ppm), doubling the value obtained with the new finish.

342 Moreover, after turning off the lamp, in box 2 where the commercial finish was applied, VOCs concentration
 343 took 3 hours to return to a steady state value, 0.45 ppm instead of the initial 0.43 ppm. On the other hand, in
 344 box 1, VOCs concentration took only 1 hour to return to a steady state value, 0.27 ppm instead of initial 0.29
 345 ppm.

346



347
 348 Figures 6 Comparison of the two boxes at the increase of temperature and increase of VOCs concentration due
 349 to T increase by irradiation. The number 1 is for box 1 (C-U) and number 2 for reference box 2 (C-L).
 350

351 **6.3. Comparison under high humidity regimes**

352 Figure 7 shows the air temperatures and RH values recorded in the two boxes by increasing the relative
 353 humidity at 75% for 8 h to simulate the humid phase of the NORDTEST used in the lab scale. Since boundary
 354 conditions were the same in box 1 and box 2, the detected change was caused only by the different response
 355 of the finish to the high levels of humidity.

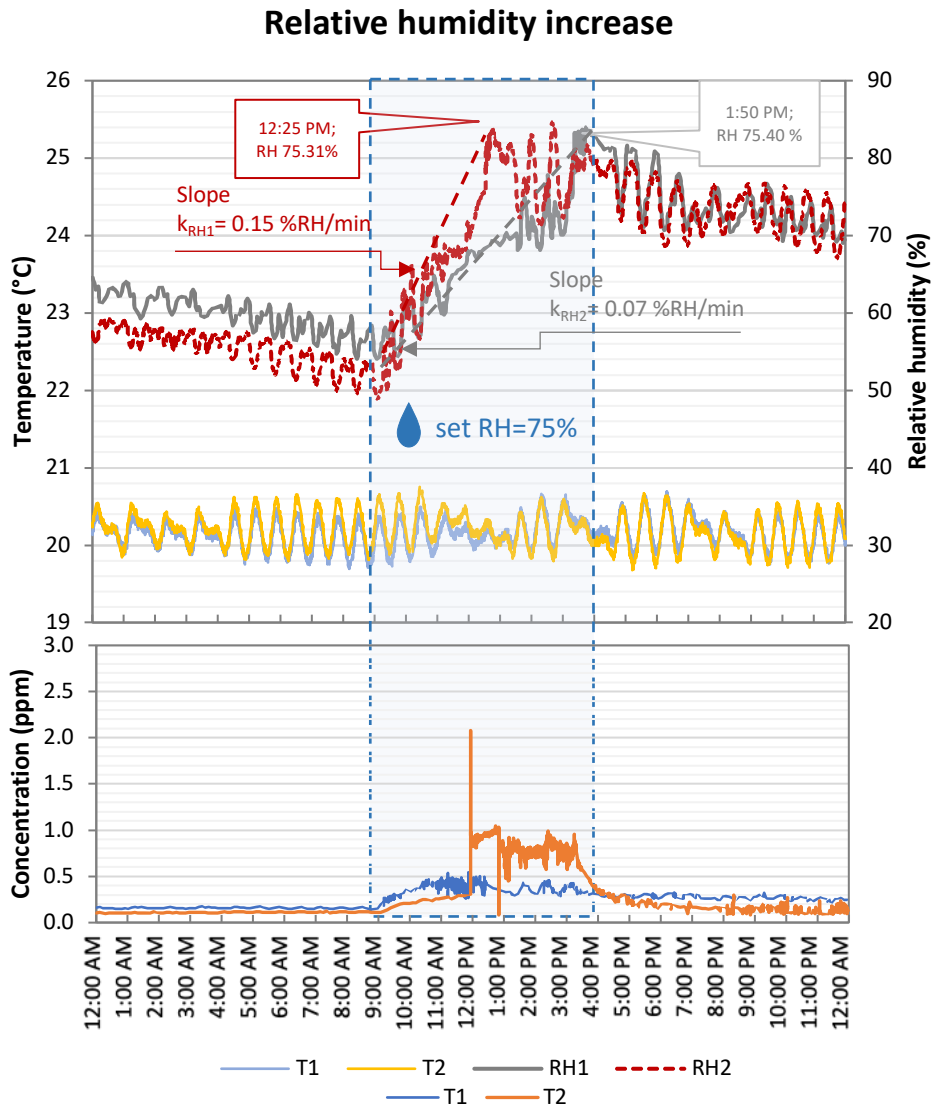
356 As soon as the humidifiers started working, the RH trend lines had different slopes: 0.15%/min for the
 357 reference C-L box 2 and 0.07%/min for C-U box 1, indicating faster RH increase in the box with the
 358 commercial products and a consequent lower moisture buffer ability. Moreover, when RH was set again at
 359 50%, RH decreased faster in box 2 than in the reference box 1, demonstrating the higher hygrometric inertia
 360 of the innovative finish compared to the reference finish. The U finish provided the environment with a greater

361 thermal and hygrometric inertia by reducing the variations of T and RH, thus confirming the better ability of
362 the innovative finish to behave as a thermal and hygrometric buffer compared to the commercial finish even
363 at a pilot scale.

364 Figure 7 shows also the VOCs concentration changes due to RH changes. However, in C-L box 2, the VOCs
365 concentration initially increased faster than in box 1, probably because MEK, adsorbed only superficially, was
366 easily removed by water. Subsequently, the VOCs concentration remained quite stable and decreased very
367 slowly due to natural decay. In case of C-U box 1, the VOCs concentration reached its maximum at a RH
368 around 70% with values much higher than in box 2, probably because in presence of high RH values water
369 saturated the porosity of the unconventional aggregates, causing the removal the high amount of MEK
370 molecules already adsorbed during the load tests (desorption process).

371 Moreover, in both cases, the increase of VOCs concentration may be related to the competition between water
372 and MEK molecules at the adsorption sites. The two spikes of VOCs concentration in box 1 and the high noise
373 in VOCs concentrations trend in both boxes could be related to a possible interference between water drops
374 generated by the humidifier and the VOCs sensor due to the relatively small volume of the box where the
375 analysis is provided. In any case, this interference did not hide the evidence of VOCs desorption discussed
376 above, demonstrating that environmental parameters such as T and RH also modify the response of finishes in
377 terms of adsorption/absorption of airborne pollutants.

378

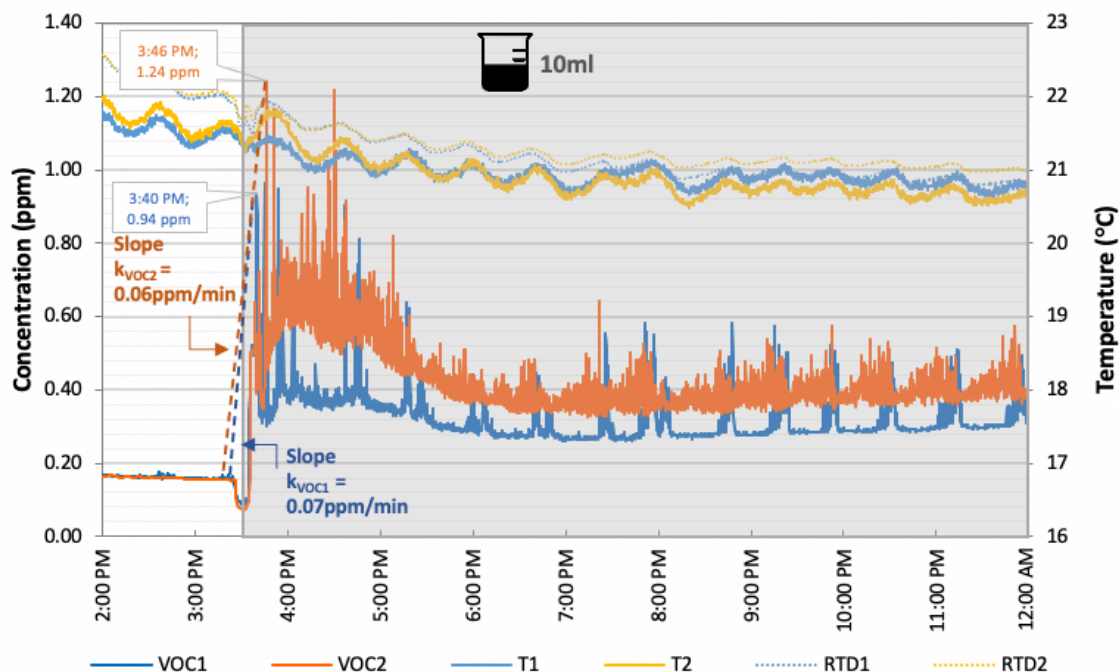


379
 380 Figure 7. Comparison of the two boxes at the increase of temperature and VOCs concentration due to RH change; the
 381 number 1 is for box 1 (C-U) and number 2 for reference box 2 (C-L).
 382

383 6.4. Comparison under high VOCs concentration

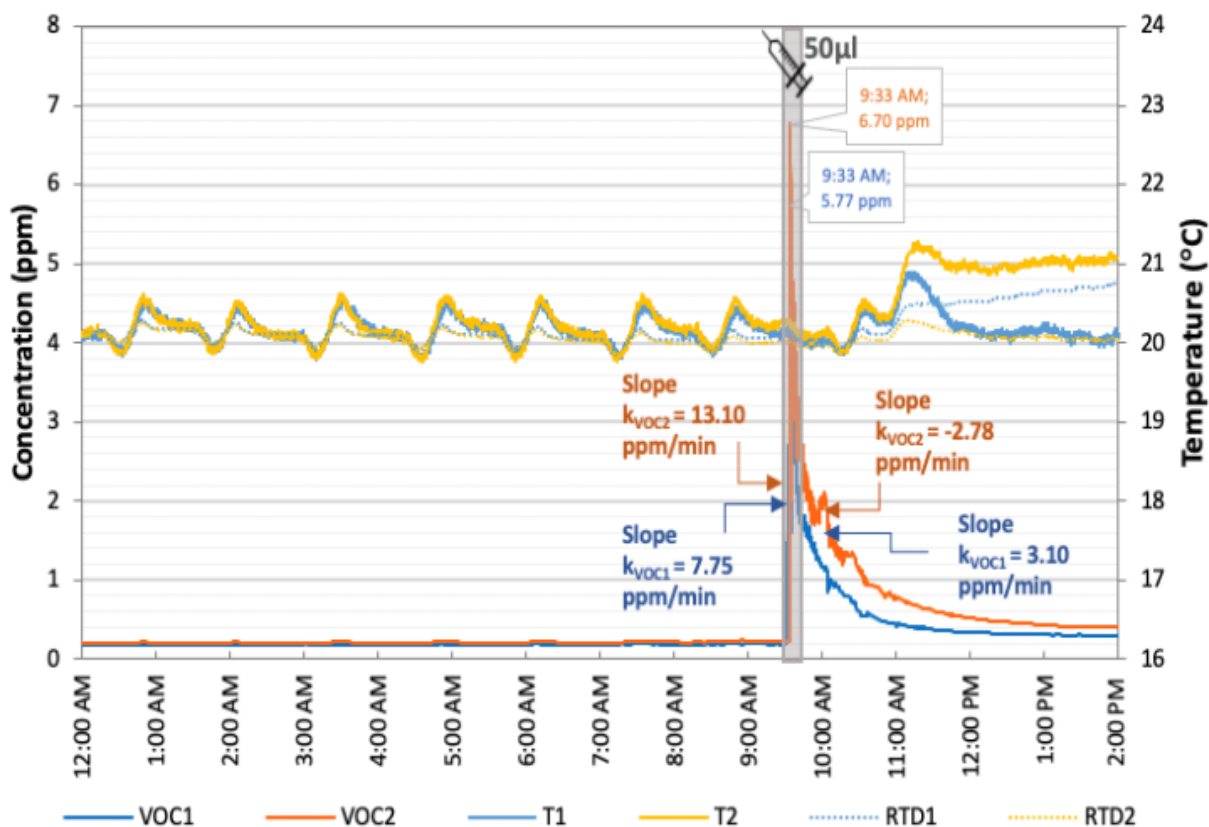
384 Figure 8 reports the environmental concentration of MEK after a static load. Before loading, in the two boxes
 385 VOCs were approximately 0.17 ppm. As soon as the MEK load a was inserted into the boxes, the VOCs
 386 sensors recorded an increase in concentration, followed by a steep slope and then a stationary phase, meaning
 387 that an equilibrium condition was reached. Although the values of the slopes of the initial phase were similar
 388 (0.06 ppm/min and 0.07 ppm/min for C-L and C-U, respectively), in the reference box 2 the MEK
 389 concentration increased up to 1.24 ppm, and then decreased very slowly in about 2.5 hours until stabilizing
 390 around approximately 0.39 pp. In box 1 with the new finish, the MEK concentration rose up to 0.94 ppm, and
 391 stabilized at 0.30 ppm, a lower value than that recorded in the reference box, confirming the depolluting ability
 392 of the innovative finish U. This demonstrates the natural concentration decay occurred in the two boxes and
 393 the passive adsorption property of the two materials, but also the greater depolluting efficacy of the innovative

394 finish U compared to L, since both the corresponding stabilization phase and the final concentration values
 395 were lower than the ones detected with L: at low concentrations, before loading, the curves coincided, but at
 396 high concentrations the U finish retained up to 20% more VOCs molecules than the reference L finish.
 397 Additionally, high concentration of VOCs implies high noise in both VOCs concentrations paths but more
 398 evident when C-L is applied since it was not able to adsorb as much MEK as C-U.
 399



400
 401 Figures 8. VOCs concentration and temperatures in the two boxes when static load of MEK was provided. The number 1
 402 is for box 1 (C-U) and number 2 for reference box 2 (C-L).
 403

404 Figure 9 shows the results obtained after the spot load of 50 μ l MEK (corresponding to a vapour concentration
 405 equal to 2.7 ppm). The load was followed by a peak in VOCs concentration, with a maximum value of 5.77
 406 ppm in box 1 and 6.70 ppm in the reference box 2. Subsequently, for the entire monitoring period, C-U
 407 maintained lower values of VOCs concentration compared to C-L in the box. The slopes of the decreasing
 408 phase of the spot load were equal to -2.78 ppm/min and -3.10 ppm/min for C-L and C-U, respectively.



409

410 Figure 9. VOCs concentration and temperatures in the two boxes with spots loads of 50 µl of MEK in the test boxes. The number 1 is

411 for box 1 (C-U) and number 2 for reference box 2 (C-L).

412

413 6.4. Comparison under high VOCs concentration and photocatalysis activation

414 Figure 10 shows the results obtained after the spot load of 100 µl MEK (corresponding to a vapour

415 concentration equal to 5.3 ppm) and the activation of TiO₂ photocatalytic function by the UV lamp. The MEK

416 load caused the occurrence of VOCs peaks in both boxes: 13.40 ppm and 4.98 ppm with a slope of 4.26

417 ppm/min for C-L and 0.52 ppm/min for C-U. Moreover, 30 mins after the spot load, the UV lamp was turned

418 on for 1.5 h. During initial lighting, the VOCs concentration assumed divergent trends: in reference box 2 the

419 L finish emitted VOCs +0.27 ppm for approximately 5 minutes, while in box 1 the U finish continued to adsorb

420 pollutants, -0.26 ppm for the same time, with a difference in MEK concentration of 0.76 ppm between the two

421 boxes. After 10 minutes of lighting, in reference box 2 the MEK concentration started to decrease again due

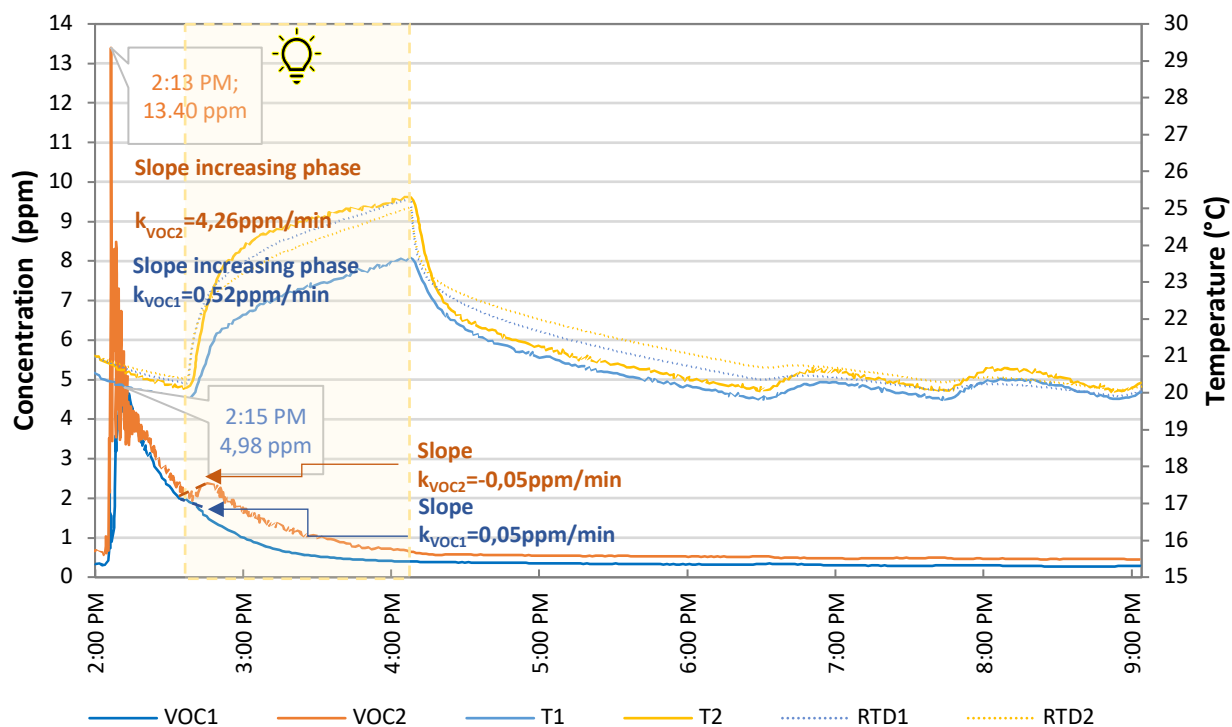
422 to the natural decay of MEK. The depollution kinetics of the two combined processes (adsorption and

423 photocatalysis) was also higher in C-U (0.05 ppm/min) than in C-L (-0.05 ppm/min). This means that in C-L

424 only desorption and natural decay of MEK occurred, while in C-U also photocatalytic activity occurred, thanks

425 to the presence of TiO₂, which decreased the VOCs concentration.

426



427

428 Figure 10. VOCs concentration and temperatures in the two boxes with spots load of 100 μ l of MEK and TiO₂ photocatalytic activation
 429 through UV lamp. The number 1 is for box 1 (C-U) and number 2 for reference box 2 (C-L).

430

431 Figure 6 shows the desorption properties of both finishes and the influence of T change on the adsorbed VOCs
 432 with no VOCs load/sources added in the environment. As previously described in section 6.2, the increase of
 433 T was provided by irradiation (UV irradiation 3 W/m² and vis irradiation 18 W/m²) with a solar lamp (300W
 434 Ultravita LUX, Osram) for 1 h, raising T from 20 °C to 28 °C. In general, data showed an increase in VOCs
 435 release under higher temperature confirming the results obtained by [14]. The increase in T increased VOCs
 436 concentration inside box 1 and box 2 of 13% (from 0.27 ppm to 0.31 ppm) and 21% (from 0.43 ppm to 0.53
 437 ppm) after one hour of irradiation, respectively; with the L finish the increase was up to double that of C-U
 438 finish.

439 Moreover, after turning off the lamp in box 2 with L finish, VOCs concentration took 3 hours to return to a
 440 steady state value, 0.45 ppm instead of the initial 0.43 ppm. On the other hand, in box 1 with U finish, VOCs
 441 concentration took only 1 hour to return to a steady state value, 0.27 ppm instead of the initial 0.29 ppm.

442

443 7. Discussion

444 The paper concerns the evaluation of thermo-hygrometric and depolluting abilities of an innovative finish (U)
 445 in both laboratory and pilot scale tests, in comparison with a commercial one (L).

446 The laboratory tests demonstrated that the finish, regardless of whether commercial or innovative, had a
 447 negligible effect on thermal conductivity and water vapour permeability due to the low thickness of the layer

448 (few millimetres). The hygroscopic buffer capacity and depolluting abilities of the new finish improved in the
449 same way, more than doubling the reference solution, since the two properties are linked to the high absorbing
450 capacities of the adopted unconventional aggregates, as demonstrated also in previous works [46] [44].

451 The on-site tests confirmed that the innovative finish can adsorb more VOCs than a reference commercial
452 finish even at a pilot scale. It can be explained by a different composition and by the inclusion of
453 unconventional aggregates that affect the final structure of the mortar, resulting in higher porosity and enabling
454 it to trap and adsorb more molecules, as reported in previously [50].

455 During the on-site tests, it was also demonstrated that building materials and furnishings can adsorb and desorb
456 (re-emit) indoor air pollutants. This phenomenon is called “the sink effect” and can involve two processes: i)
457 the adsorption (and desorption) of VOCs on a material surface, and ii) the diffusion of VOCs from the surface
458 into the material interior. On the other hand, if the sink effect is 100% reversible, meaning that adsorbed VOCs
459 molecules are later released, it might also lead to a protracted low-level exposure of building occupants to
460 VOCs. Good adsorption capacity, even if irreversible, and a very low desorption rate, if reversible, are
461 characteristics of an ideal building material for improving IAQ. This ensures that the concentration of emitted
462 VOCs is considerably below both the perceived threshold and the IAQ requirements.

463 The moisture buffering ability of a finish is important [56] not only for occupants’ health, but also for energy
464 saving issues. In particular, the efficacy of moisture buffering materials is recognized for temperate zones with
465 high humidity differences between day and night [35] [57].

466 The test demonstrated that environmental parameters such as T and RH modify the response of finishes in
467 terms of adsorption/absorption of airborne pollutants, with the innovative finish more influenced by RH,
468 whereas the commercial one more influenced by T.

469 Data showed an increase in VOCs release under higher T confirming the results obtained by [14]. Moreover,
470 when RH was reduced after the high regime of 75%, the VOCs concentrations decreased in the two boxes,
471 demonstrating the desorption of MEK from the superficial open porosity (active sites) of both finishes
472 occurring at rising RH [58] [59]. However, between the two materials, the C-U system showed higher VOCs
473 depolluting capacity; this was attributed both to the presence of adsorbent materials [60] and to the presence
474 of TiO₂ in the mix-design since TiO₂ has already been proven to be able to mineralize pollutants [61].

475

476 **8. Conclusions**

477 The present work was aimed at verifying the efficacy of an innovative finish to improve IAQ in both laboratory
478 and pilot scale tests in comparison with a commercial finish for the same applications. In particular, tests were
479 conducted to evaluate the ability of the two finishes to passively improve indoor comfort and health in terms
480 of thermo-hygroscopic properties such as temperature and moisture buffering capacity and and depolluting
481 abilities such as Volatile Organic Compounds (VOCs) adsorption and desorption.

482 From the laboratory to pilot scale results the following main conclusions can be drawn:

- 483 - Even if applied at very small thickness, the innovative finish is able to reduce the
484 temperature peaks (about -1.5°C) better than the commercial finish at pilot scale.

- 485 - The innovative finish has a higher moisture buffering ability than the commercial finish both
486 at lab and pilot scales.
- 487 - The indoor VOCs concentration is influenced by both T and RH and the increase of these
488 parameters causes VOCs desorption.
- 489 - Increase of T causes VOCs desorption but at pilot scale the tested T of 25 °C does not cause
490 the release of adsorbed VOCs from the active sites of the unconventional aggregates.
- 491 - Increase in RH causes a VOCs removal from the internal porosity of the adsorbent
492 aggregates. For this reason, the application of the new finish in indoor environment with
493 high RH values and RH variations, should be carefully investigated.
- 494 - Between the two finishes, the innovative one showed a higher VOCs depolluting capacity
495 both at lab and pilot scale. This is due not only to photooxidation by TiO₂ able to mineralize
496 the pollutants if activated by UV light but also to VOCs adsorption by unconventional
497 aggregates of the patented mix.
- 498

499 **Patent**

500 *Italian Patent 102017000033750*

501

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510

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